

## Regional and global perspectives of honey as a record of lead in the environment

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1 **Regional and global perspectives of honey as a record of lead in the environment**

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4 **44 ABSTRACT**

5 45 Honey from *Apis mellifera* is a useful and inexpensive biomonitor for mapping metal  
6 46 distributions in urban centers. The sampling resolution of a biomonitoring survey (e.g., city  
7 47 versus global scale) determines which geochemical processes are reflected in the results. This  
8 48 study presents Pb isotopic compositions and metal concentrations in honey from around the  
9 49 world, sampled at varying resolutions: honey from Canada (n = 21), the United States (n = 111),  
10 50 Belgium (n = 25), and New Zealand (n = 10), with additional samples from Afghanistan, Brazil,  
11 51 Cuba, Germany, Liberia, Taiwan, and Turkey. Honey was sampled at high resolution in two  
12 52 uniquely different land-use settings (New York Metro Area and the Hawaiian island of Kaua'i),  
13 53 at regional-scale resolution in eastern North America (including the Great Lakes region), and Pb  
14 54 isotopic compositions of all samples were compared on a global scale. At high sampling  
15 55 resolution, metal concentrations in honey reveal spatially significant concentration gradients: in  
16 56 New York City, metals associated with human activity and city infrastructure (e.g., Pb, Sb, Ti, V)  
17 57 are more concentrated in honey collected within the city compared to honey from upstate New  
18 58 York, and metal concentrations in honey from Kaua'i suggest polluting effects of nearby  
19 59 agricultural operations. At lower resolution (regional and global scales), lead isotopic  
20 60 compositions of honey are more useful than metal concentrations in revealing large-scale Pb  
21 61 processes (e.g., the enduring legacy of global leaded gasoline use throughout the twentieth  
22 62 century) and the continental origin of the honey. Lead isotopic compositions of honey collected  
23 63 from N. America (especially from the eastern USA) are more radiogenic ( $^{206}\text{Pb}/^{207}\text{Pb}$ : 1.132-  
24 64 1.253,  $^{208}\text{Pb}/^{206}\text{Pb}$ : 2.001-2.129) compared to European honey, and honey from New Zealand,  
25 65 which has the least radiogenic isotopic compositions measured in this study ( $^{206}\text{Pb}/^{207}\text{Pb}$ : 1.077-  
26 66 1.160,  $^{208}\text{Pb}/^{206}\text{Pb}$ : 2.090-2.187). Thus, biomonitoring using honey at different resolutions  
27 67 reflects differing processes and, to some extent, a honey *terroir* defined by the Pb isotopic  
28 68 composition. The data presented here provide important (and current) global context for future  
29 69 studies that utilize Pb isotopes in honey. Moreover, this study exhibits community science in  
30 70 action, as most of the honey was collected by collaborators around the world, working directly  
31 71 with local apiarists and hobby beekeepers.  
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33 **73 KEYWORDS**

34 74 honey, lead isotopes, New York City, Kaua'i, global lead array  
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93 BeeHIVE Research Excellence Cluster. Additional support for K. Smith was provided by UBC via  
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## 1. INTRODUCTION

Since the Industrial Revolution, natural emissions of certain metals and metalloids (e.g., Pb, V, Sb, Hg, Ni, Cd, Zn) have been dwarfed by anthropogenic emissions (Han et al., 2002; Nriagu, 1989; Rauch and Pacyna, 2009). Of ongoing concern is lead (Pb), a neurotoxic metal and known teratogen, which is difficult (and costly) to remediate once deposited in the environment (Landrigan et al., 2018; Mielke et al., 2011; O'Connor et al., 2020). Natural sources of Pb in the environment include volcanic emissions, sea spray, and weathering and transport of local rocks, soils, and sediments. Anthropogenic inputs include emissions from mining operations, smelting, fuel combustion (low-grade coal, fuel oil, gasoline), and battery manufacturing and recycling. The most notable modern anthropogenic inputs of Pb to the environment are the use of leaded gasoline throughout most of the twentieth century and leaded paints resulting in the contamination of indoor dust and residential topsoil (affecting gardens and playgrounds) (Filippelli and Laidlaw, 2010; Mielke, 2018; Mielke and Reagan, 1998; O'Connor et al., 2018). The chemical 'legacy' of leaded gasolines and paints govern Pb-focused environmental studies to this day, decades after termination of their widespread use, including soil remediation efforts and public health concerns relating to elevated blood lead levels, especially in children (Filippelli et al., 2015; Zahran et al., 2013).

The results of modern lead use are apparent on local and global scales. From point sources, many of which release coarse particulates ( $> 10 \mu\text{m}$ ), the extent of environmental enrichment for Pb is local or regional since coarse particulates have shorter atmospheric residence time, minimizing transport and causing an exponential decrease in pollutant concentration in topsoils as a function of distance from the source (Dong et al., 2020; Masri et al., 2015; Reimann et al., 2011). For example, degradation of exterior paints, resuspension of road dust or contaminated soil, or mechanical wear from metal refining processes all tend to cause local Pb pollution (Laidlaw and Filippelli, 2008; Weiss et al., 2006). Conversely, fine particulates ( $< 2.5 \mu\text{m}$ ) have longer residence times in the atmosphere before settling/deposition, so they are more susceptible to transport than coarse particulates (Wilson et al., 2005). The consequences of leaded gasoline use are evident on local, continental, hemispheric, and global scales because combustion of leaded gasoline produces both coarse particulates that are deposited along roadways, especially in areas of high traffic (i.e., cities),

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162 and fine particulates, approximately 35 % of which fall into the ‘ultrafine’ size fraction (< 0.1  
163  $\mu\text{m}$ ) (Mielke et al., 2011) that are subject to long-range (i.e., transcontinental) transport. In  
164 general, high-resolution geochemical mapping reveals metal distribution at smaller scales,  
165 highlighting the impact of local point or diffuse sources, which tend to be more anthropogenic  
166 (e.g., Pb from an urban center, a smelter, or a mining operation) (Demetriades et al., 2010;  
167 Locutura and Bel-lan, 2011; Mielke, 1994). Lower-resolution studies reveal continental and  
168 global-scale processes, most of which are natural (i.e., crustal geochemistry, glaciation  
169 boundaries, large-scale climate trends) (Smith et al., 2018), but also reflect some anthropogenic  
170 activity, of which the global use of leaded gasoline is a key component.

171 The extent and history of Pb pollution is recorded in natural environmental archives and  
172 the analysis of Pb stable isotopes ( $^{208}\text{Pb}$ ,  $^{207}\text{Pb}$ ,  $^{206}\text{Pb}$ ,  $^{204}\text{Pb}$ ) in these archives aids researchers in  
173 differentiating Pb sources (natural, anthropogenic, and legacy anthropogenic) (e.g., Komárek et  
174 al., 2008) on various scales. Environmental proxies for Pb with local or regional spatiotemporal  
175 constraints include lichens and mosses (Simonetti et al., 2003), fish and shellfish (Li et al., 2020;  
176 Shiel et al., 2012), and tree rings (Marcantonio et al., 1998; Novak et al., 2010; Patrick and  
177 Farmer, 2006). Sediment, peat, ice, and snow cores provide anywhere from continental to  
178 global-scale records of Pb pollution (e.g., Hong et al., 1994; McConnell et al., 2018; Shotyky et al.,  
179 1998; Weiss et al., 1999). Agricultural and consumer products (e.g., wine, vinegar, tobacco)  
180 have Pb isotopic compositions that reflect their local environment of origin, thus serving as an  
181 ‘archive’ of the Pb chemistry where the product originated (Epova et al., 2020; Guo et al., 2015;  
182 Kristensen et al., 2016; Medina et al., 2000; Ndung’u et al., 2011).

183 Honey, and European honeybees (*Apis mellifera*), are effective biomonitors of chemicals  
184 in the environment. Honey and bees have been used for monitoring pesticide residues on or  
185 near farming operations (e.g., Codling et al., 2016; de Oliveira et al., 2016; Porrini et al., 2003)  
186 and for determining metal distributions (including Pb) near point sources of metal pollution  
187 (Smith et al., 2019; Zhou et al., 2018b). Since bees forage within a few kilometers of their hive  
188 (Eckert, 1933) and passively collect dust and particulates while they forage for pollen and  
189 nectar, the bees and their products (honey, bee pollen, wax) develop a chemical composition of  
190 spatial significance, that reflects their foraging environment (Negri et al., 2015; Smith and Weis,

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4 191 2020; Taylor, 2019). City-scale investigations (tens of kilometers) of metal distributions using  
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6 192 honey and hive products are numerous (e.g., Bromenshenk et al., 1985; Conti and Botrè, 2001;  
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8 193 Giglio et al., 2017; Skorbiłowicz et al., 2018; Van der Steen et al., 2015; Zarić et al., 2016).  
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10 194 However, source apportionment investigations using Pb isotopic compositions in honey (and  
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12 195 other hive products) are still somewhat rare, and the existing studies are limited to city-scale  
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14 196 and point-source investigations (Smith et al., 2019; Smith and Weis, 2020; Zhou et al., 2018b).  
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16 197 Published metal concentration data compiled for honey from around the world is typically  
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18 198 reported for food science and food quality purposes (e.g. Solayman et al. 2016), rather than for  
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20 199 geochemical surveys, or verification of honey authenticity, which are sometimes used in  
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22 200 conjunction with carbon stable isotopes or strontium isotopes (Baroni et al., 2015; Batista et al.,  
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24 201 2012; Zhou et al., 2018c). Missing from the literature is a global Pb isotopic database of honey.  
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26 202 For honey to be a viable biomonitor of Pb in the future, it is essential to provide a current,  
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28 203 global context for existing and future city-scale studies.

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30 204 In this study, we present applications of honey as a biomonitor for Pb in two disparate  
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32 205 regions: the very populated and industrialized eastern North America (n = 58, featuring samples  
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34 206 from the New York Metropolitan Area and the Great Lakes region), and the very remote  
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36 207 Hawaiian Islands (n = 52). We then examine honey as a biomonitor for Pb on a global scale,  
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38 208 using a suite of honey samples collected around the world (n = 181). Using these data, we  
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40 209 demonstrate how honey fits within the modern global Pb array, as defined by previous regional  
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42 210 and city-scale honey studies as well as anthropogenic and natural Pb isotopic compositions  
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44 211 measured in consumer products and aerosols. The database for Pb isotopes in honey presented  
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46 212 here validates a ‘community science’ approach, as many of the samples were sourced from  
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48 213 collaborators and hobby apiarists around the world.

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## 50 51 215 **2. MATERIALS AND METHODS**

### 52 53 216 **2.1. Field and sampling information**

#### 54 55 217 *2.1.1. Field methods*

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57 218 Honey samples were collected by community scientists who volunteered honey from their  
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59 219 backyard apiaries or were purchased (as commercially available honey) in the country of origin

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220 (i.e., while traveling) or in Canada and the United States (USA) as imported products.

221 Community scientists used clean sampling vials, sampling instructions, and field worksheets for  
222 consistent metadata collection, after Smith and Weis (2020).

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224 *2.1.2. Sampling sites*

225 Eastern North America (including the Great Lakes region) hosts the majority of the continent’s  
226 population and is rich in large urban centers, which feature concentrated areas of industrial  
227 activity (U.S. EPA, 2017; Canada NPRI, 2017). The New York Metro Area includes the five  
228 boroughs of New York City (NYC) and extends into parts of New Jersey (NJ), Pennsylvania (PA),  
229 and Connecticut (CT) and is home to more than 23 million people (U.S. Census Bureau, 2016). A  
230 community science effort, notably led by a high school student, provided honey samples from  
231 NY Metro Area (n = 29) and samples from upstate NY, PA, and CT (n = 14), collected during  
232 summer 2019.

233         The Hawaiian Islands offer a stark contrast to eastern North America. With a total  
234 population of around 1.4 million, there are few urban and industrial (i.e., manufacturing)  
235 centers on the Hawaiian Islands; the main industries are agriculture, tourism, and military  
236 activities. The Hawaiian Islands are very remote (a volcanic island chain, located in the central  
237 Pacific Ocean), and each island is small (< 10,500 km<sup>2</sup>), with a compact watershed and wet,  
238 tropical climate. These factors, and year-round crop production, make the Hawaiian Islands a  
239 unique setting to investigate honey as a biomonitor. Honey was sampled on the Island of Kaua’i  
240 (31 samples from 22 unique sites) through community sampling efforts between 2013-2016  
241 (Berg et al., 2018). These Kaua’i samples were collected directly from feral hives, managed  
242 hives, or as commercial honey from local producers. For context, additional commercial honey  
243 samples were purchased on the Islands of O’ahu, Maui, and Hawai’i (Big Island) in 2017-2018.  
244 Other significant community sampling efforts for this study occurred in and around Brussels,  
245 Belgium; Madison, Wisconsin (WI, USA); and in New Zealand (from North and South Island).

246         Mississippi Valley Type (MVT) galena samples (n = 4) from Lafayette County, WI, part of  
247 the Upper Mississippi Valley Mining District, were provided by the WI Geological and Natural  
248 History Survey. While not used in tetraethyl lead (TEL) production to the same extent as the  
249 southeast (SE MO) MVT ores, these samples provide important geochemical context for Upper



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4 250 Mississippi Valley Pb ores used throughout the Great Lakes Megalopolis Manufacturing Belt  
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6 251 during and after the Industrial Revolution.  
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10 253 **2.2. Sample preparation and analysis**

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12 254 *2.2.1. Honey*

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14 255 Samples were prepared and analyzed in clean laboratories at the Pacific Centre for Isotopic and  
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16 256 Geochemical Research (PCIGR) at the University of British Columbia (UBC), except for the honey  
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18 257 and galena collected in Wisconsin (described in section 2.2.2). All acids reagents used at PCIGR  
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20 258 were purchased as highest-purity grade available or were distilled in-house, and only ultrapure  
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22 259 H<sub>2</sub>O was used (18.2 MΩ·cm). Honey samples were digested and analyzed using the methods  
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24 260 described by Smith et al. (2019): microwave digested in concentrated HNO<sub>3</sub>, then evaporated  
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26 261 and reconstituted in 2 % HNO<sub>3</sub> + 10 ng/g indium (In, added as an internal standard for metals  
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28 262 analysis by inductively-coupled plasma mass spectrometry, ICPMS). Each honey digestion batch  
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30 263 also included the standard reference material (SRM) NIST 1568b (Rice Flour, National Institute  
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32 264 of Standards and Technology, Gaithersburg, MD, USA).

33 265 Metal concentrations (Mg, Al, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, Ga, As, Rb, Sr, Zr, Mo, Cd,  
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35 266 Sn, Sb, Ba, Pb) were determined by ICPMS (Agilent 7700x, Agilent Technologies, Santa Clara, CA,  
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37 267 USA). These elements include major and trace elements that originate from lithogenic and  
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39 268 anthropogenic sources (or both), are measurable in honey, and have been useful in past studies  
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41 269 for city-scale metal distribution surveys using honey (Smith et al., 2019; Smith and Weis, 2020).  
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43 270 Lead isotopic compositions were measured in the same digests by high resolution (HR-) ICPMS  
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45 271 (Nu AttoM, Nu Instruments Ltd., Wrexham, UK). In-house multicollector (MC-) ICPMS  
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47 272 verification of NIST 1568b was previously presented by Smith and Weis (2020).  
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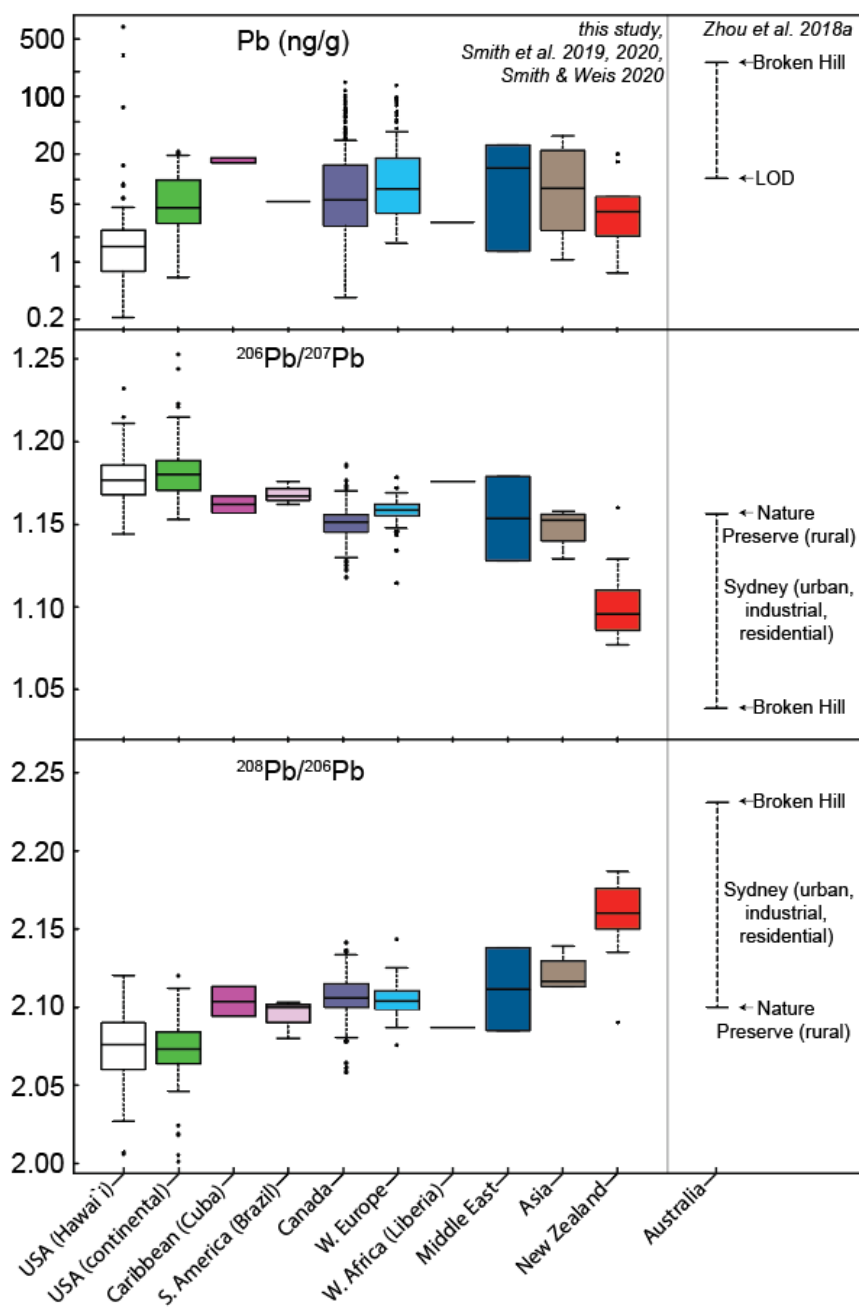
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51 274 *2.2.2. Wisconsin honey and galena*

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53 275 The Wisconsin honey (n = 7) and galena (n = 4) samples were prepared and analyzed at the  
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55 276 Trace Element Clean Lab (TECL) facility at the Wisconsin State Laboratory of Hygiene, University  
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57 277 of Wisconsin-Madison. All reagents were of Optima-grade, purchased from Thermo Fisher  
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59 278 Scientific (Waltham, MA, USA). Honey samples were weighed into pre-cleaned 30 mL Teflon  
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4 279 beakers (Savillex Corporation, MN, USA) and digested using 10 mL of 16 M HNO<sub>3</sub>. Samples were  
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6 280 dried, then re-digested in 9.9 mL 16 M HNO<sub>3</sub> + 0.1 mL 30% H<sub>2</sub>O<sub>2</sub>. One mL of this digest was  
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8 281 removed and diluted in 2% HNO<sub>3</sub> for metals analysis using an Element2 HR-ICPMS (Thermo  
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10 282 Fisher Scientific). The remaining solution was dried and reconstituted in 5 mL of 1 M HBr.  
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12 283 Samples were loaded onto BioRad PolyPrep columns (BioRad Laboratories, Inc., CA, USA)  
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14 284 containing 1 mL of pre-cleaned AG1-X8 anion exchange resin (BioRad) preconditioned in 1 M  
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16 285 HBr. Matrix elements were removed using washes of 1 M HBr, and the Pb fraction was  
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18 286 collected using ultrapure H<sub>2</sub>O and 1 M HNO<sub>3</sub>. Galena samples were dissolved using 16 M HNO<sub>3</sub>  
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20 287 and were not purified prior to analysis. Final samples were diluted in 2 % HNO<sub>3</sub> and spiked with  
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22 288 Tl to correct for mass bias. Pb and Tl isotopes were analyzed using a Neptune Plus MC-ICPMS  
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24 289 (Thermo Fisher Scientific) in static mode using 6 faraday collectors, with an additional collector  
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26 290 used to monitor for Hg interferences. Honey sample solutions were introduced into the plasma  
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28 291 using an Aridus 3 desolvating nebulizer (Teledyne CETAC Technologies, NE, USA) and the Jet  
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30 292 sampler and X-skimmer cone configuration. Galena sample solutions were introduced into the  
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32 293 plasma using the standard cyclonic spray chamber with the standard sampler cone and X-  
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34 294 skimmer cone configuration. Lead isotopic standard NIST 981 was analyzed periodically, and  
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36 295 final isotope ratios are reported relative to NIST 981 values determined by Galer and  
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38 296 Abouchami (1998). NIST 1568b was included for quality control.

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41 298 **2.4. Calculations and statistical analysis**

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43 299 All statistical calculations were made using R, v. 4.0.2 (R Core Team, 2020). Honey data from  
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45 300 different global regions (e.g., Fig. 1) were compared using a Kruskal-Wallis rank sum test  
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47 301 (Kruskal and Wallis, 1952). If the results were significant, indicated by an adjusted  $p$ -value ( $p_{adj.}$ )  
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49 302  $< 0.05$ , then a post hoc Dunn's multiple comparison test (Dunn, 1964) was used to identify  
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51 303 significantly different subgroups.



**Figure 1.** Box plots comparing Pb concentrations and isotopic compositions for honey, by region. The plotted data include this study (Table S1, S2), and other honey data from France (included in the 'W. Europe' region, Smith et al., 2020) and previous honey data reported for Canada (Smith et al., 2019; Smith and Weis, 2020). For comparison, Australian honey data from Zhou et al. (2018b) is also included; collected from Sydney (urban, industrial, and residential area), a nature preserve (rural), and from near the Broken Hill mine (LOD = limit of detection reported by Zhou et al., 2018b). Note that there is a logarithmic scale for the Pb concentration panel (top panel), only. A summary for pairwise statistical comparison, by region (for all regions with  $\geq 10$  measurements), for metal concentrations and Pb isotopes is included in the Supplementary Material (Table S3).

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315 **3. RESULTS**

316 **3.1. Blanks and standard reference materials**

317 All metal concentrations, Pb isotope data, and summary for statistical tests are reported in the  
318 Supplementary Material (Tables S1-S3). Trace element results for microwave digestion blanks  
319 (honey digestion method) are reported in previous studies (Smith et al. 2019, Smith and Weis  
320 2020, with an average of 36 pg total Pb). Refer to Smith and Weis (2020) for in-house HR-ICPMS  
321 and MC-ICPMS results summaries for NIST 1568b (dataset doi:10.5683/SP2/Y9MKTN). An  
322 aliquot of the same NIST 1568b used at the PCIGR (UBC) was analyzed at the TECL (UW) facility  
323 (n = 2). Average MC-ICPMS results between the two facilities are in good agreement, with  
324 relative mean differences of 3.84 per mille (‰), 0.19 ‰, and 2.75 ‰ for <sup>208</sup>Pb/<sup>204</sup>Pb,  
325 <sup>207</sup>Pb/<sup>204</sup>Pb, and <sup>206</sup>Pb/<sup>204</sup>Pb, respectively. It is worthwhile to note that analytical results of  
326 aliquots from a newer bottle of NIST 1568b (purchased in 2020) were not in good agreement  
327 with previous results (from either facility), indicating that this SRM does not have a  
328 homogenous Pb isotopic composition. This supports previous claims (e.g., Pohl et al., 2017;  
329 Smith and Weis, 2020) that a suitable (well characterized and widely available) SRM would be  
330 very useful for studies using hive products for metal biomonitoring purposes.

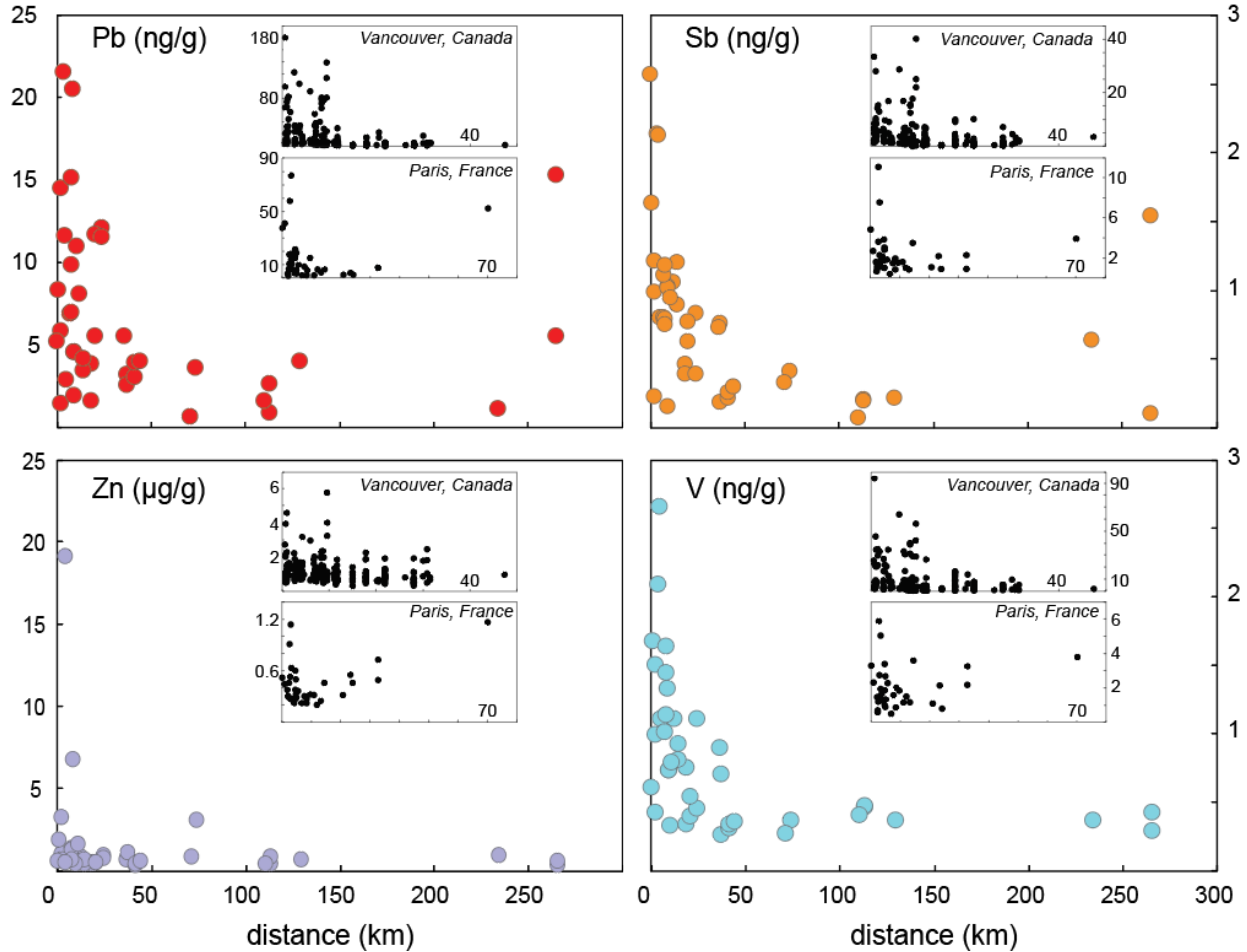
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332 **3.2. Metal concentrations**

333 Metal concentrations in all honey analyzed in this study are comparable to ranges reported in  
334 honey from all over the world, compiled by Solayman et al. (2016), and do not indicate any  
335 concern for consumer health. Our results are also generally comparable to more recent  
336 analyses of global honey by Zhou et al. (2018c), with some honey from this study exceeding  
337 their reported ranges: Kaua’i, HI (Al, Ba, Fe), Metro New York (Fe), Brazil (Cu, Mg), Belgium (Rb),  
338 and New Zealand (Rb). While the Pb concentration range for Pb in this study is large (0.2 to 709  
339 ng/g), most of the honey samples (> 98 %, 179 out of 181) have Pb contents below 135 ng/g.  
340 The suite of Hawaiian honey samples is unique in that it includes honey with the lowest (0.2  
341 ng/g, Hawai’i Is.) and highest (709 ng/g, Kaua’i) concentrations of Pb measured in this study.  
342 The two highest Pb concentrations measured in Kaua’i (709 and 317 ng/g Pb) exceed  
343 concentrations previously measured in honey from urban centers: Vancouver, BC; Broken Hill,

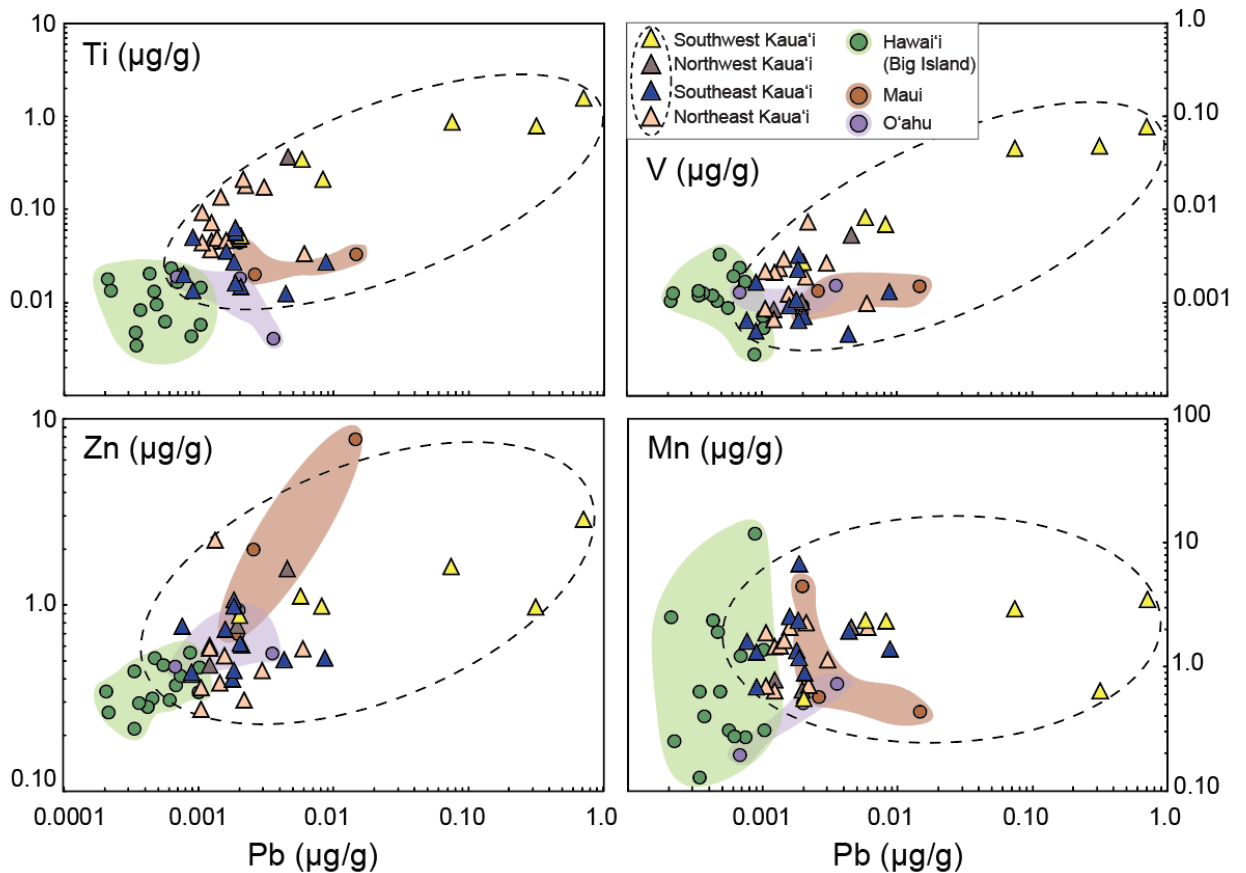
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344 Australia; Paris, France (Smith et al., 2020; Smith and Weis, 2020; Zhou et al., 2018b), but are  
345 lower than Pb concentrations reported in honey in other studies from Europe, North Africa,  
346 Malaysia, and the USA (Soleyman et al., 2016 and references therein).

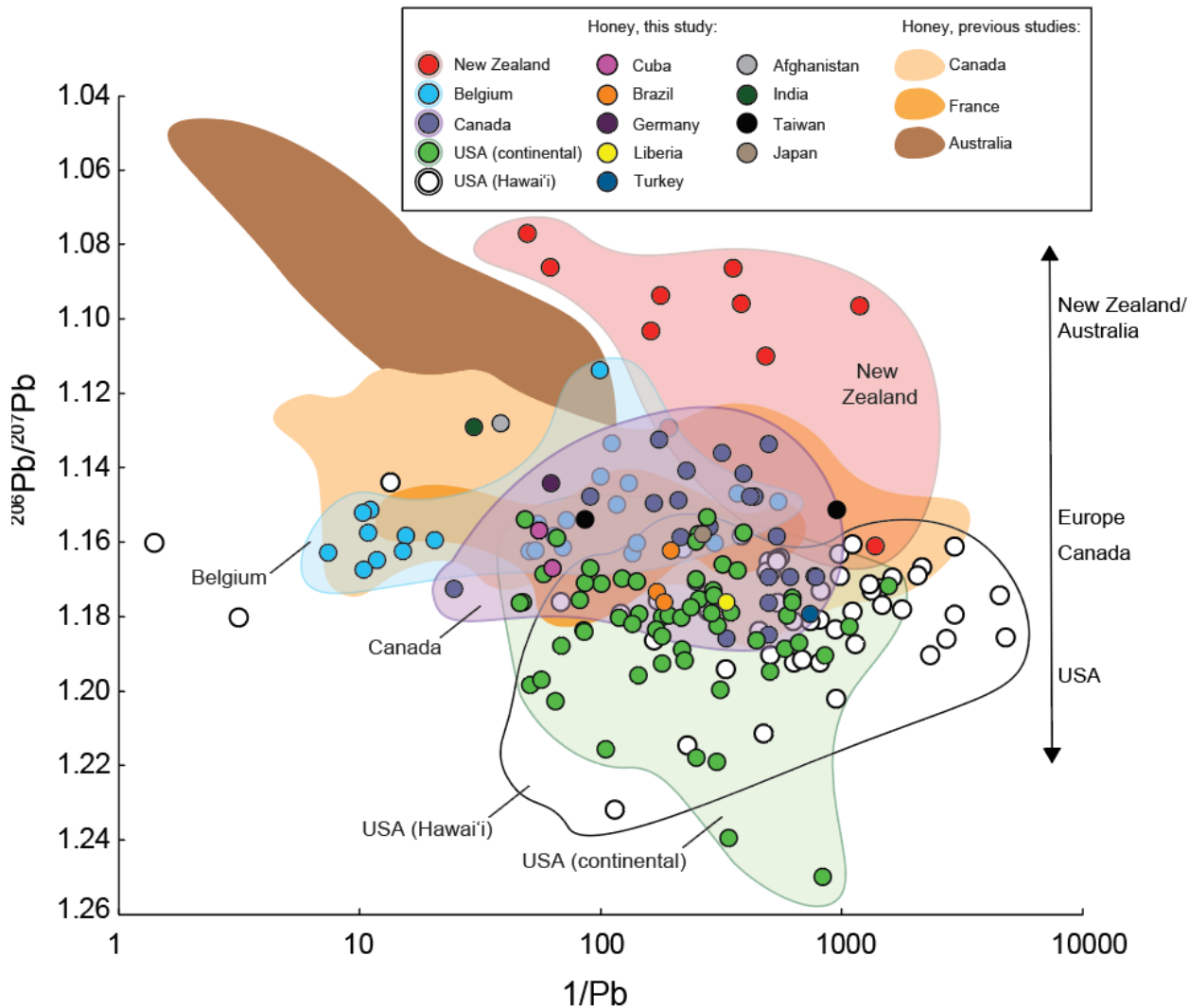


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348 **Figure 2.** Metal concentrations in honey from the NY Metro Area as a function of radial distance  
349 (km) from the center of Bryant Park, located in Midtown Manhattan. That site was selected to  
350 demonstrate the effect on certain metal concentrations in honey as a function of distance from  
351 the city center. Note that units for zinc (Zn) are  $\mu\text{g/g}$ , while others are ng/g. Concentration  
352 maps for these metals in NY Metro honey are in the Supplementary Material (Fig. S1). For figure  
353 clarity, error bars are not included (errors are reported in Table S1). For comparison, insets  
354 show concentration plots for these metals in honey from two other cities: Metro Vancouver,  
355 Canada (data compiled from Smith et al., 2019 and Smith and Weis, 2020) and Paris, France  
356 (data from Smith et al., 2020). Distances are measured from the Port of Vancouver (after Smith  
357 et al. 2019) in Metro Vancouver and from the Notre Dame cathedral in Paris. Insets have the  
358 same units as their corresponding main plot for both axes.

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 4 360 For all metal concentrations, there is no systematic, geographic variation on a regional  
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 6 361 or global scale, (Table S3, Fig. 1), but in areas where sampling density was high (high sampling  
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 8 362 resolution), the local geospatial variations in metal concentrations are apparent. For example,  
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 10 363 in New York City, we observe decreasing concentrations of certain metals in honey (Pb, Ti, V,  
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 12 364 Sb, Cd; i.e., metals often associated with anthropogenic activity) as a function of distance from  
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 14 365 central Manhattan: the highest concentrations are observed in honey from within the city (Fig.  
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 16 366 2, S1). On Kaua'i, honey on the south side of the island has uniquely elevated levels of certain  
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 18 367 metals (e.g., Pb, Cd, Ti, V, Zn, Mn) relative to honey from elsewhere on Kaua'i and from other  
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 20 368 Hawaiian Islands (Fig. 3, S2).  
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 53 371 **Figure 3.** Metal concentration bivariate plots (Ti, V, Zn, Mn vs. Pb) for honey from four Hawaiian  
 54 372 Islands: Hawai'i, Maui, O'ahu, and Kaua'i. All units are µg/g; note the logarithmic scale. For  
 55 373 figure clarity, error bars are not included (errors are reported in Table S1).  
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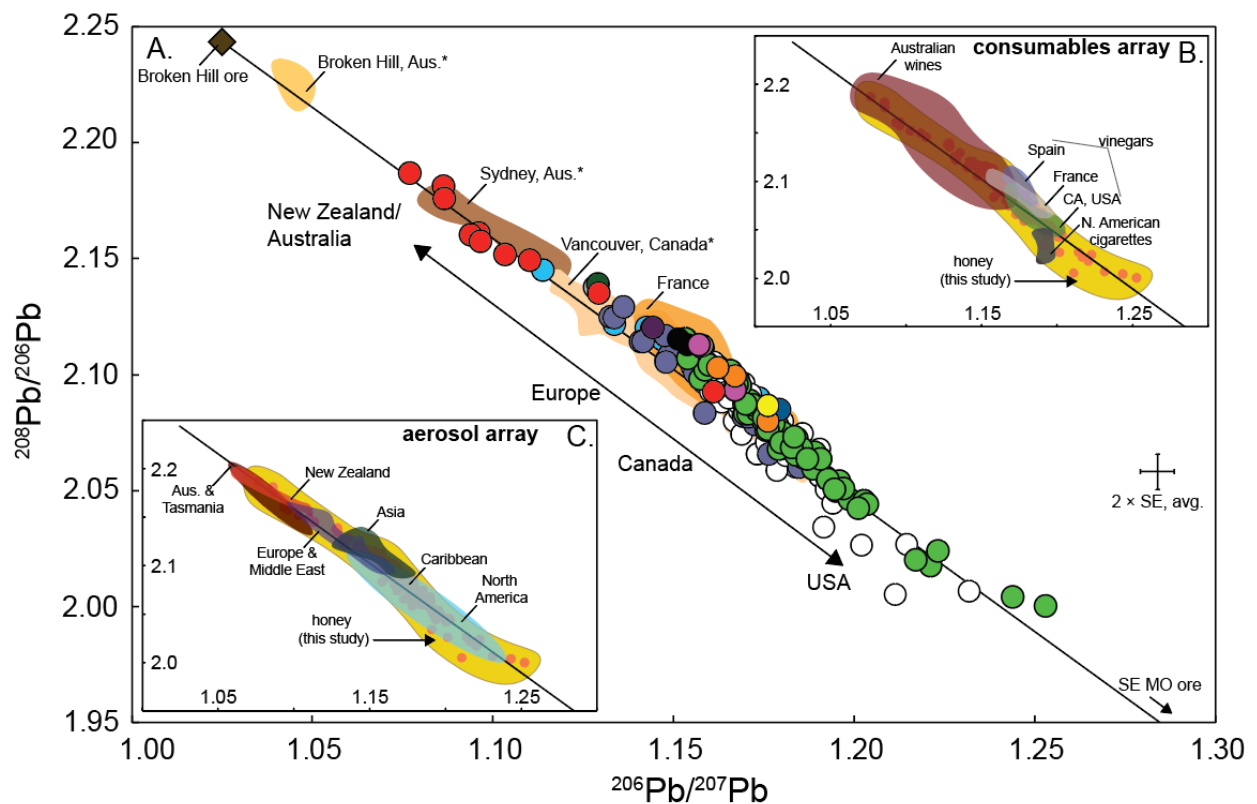
**Figure 4.** A plot of  $1/\text{Pb}$  ( $\text{g}/\mu\text{g}$ ) versus  $^{206}\text{Pb}/^{207}\text{Pb}$  for all honey analyzed in this study. Honey from previous studies is included for context (as shaded fields): Broken Hill and Sydney, Australia (Zhou et al., 2018b); France (Smith et al., 2020); Vancouver, Canada (Smith et al., 2019; Smith and Weis, 2020). Regions from this study that include  $\geq 10$  samples are also shaded (Continental USA and Hawai'i, Canada, Belgium, New Zealand).

### 3.3 Pb isotopes

All honey samples analyzed in this study have a Pb isotopic range of 1.077 to 1.253 and 2.001-2.187 for  $^{206}\text{Pb}/^{207}\text{Pb}$  and  $^{208}\text{Pb}/^{206}\text{Pb}$ , respectively (Table S1-S2). The Pb isotopic compositions of the honey vary based on their region of origin. In general, honey collected from N. America (especially from the eastern USA) is significantly more radiogenic (higher  $^{206}\text{Pb}/^{207}\text{Pb}$ , lower  $^{208}\text{Pb}/^{206}\text{Pb}$ ) compared to European honey, and compared to honey from New Zealand, which

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has the least radiogenic isotopic compositions measured in this study ( $p_{adj} < 0.05$  for  $^{206}\text{Pb}/^{207}\text{Pb}$  and  $^{208}\text{Pb}/^{206}\text{Pb}$ , Fig. 1, 4). Based on geology and histories of anthropogenic Pb use (e.g., which Pb gasoline additives were used in various parts of the world), the isotopic results were as expected for all geographic regions given that they plot along the global Pb array, which is strongly governed by the twentieth century's leaded gasoline mixing line (Fig. 5). We observe one exception in this study: honey collected from the northeast part of Kaua'i deviates slightly from the Pb isotopic compositions of other Hawaiian honey (and the global Pb array), Hawaiian geology, and input from Eurasian dust (Fig. 6).

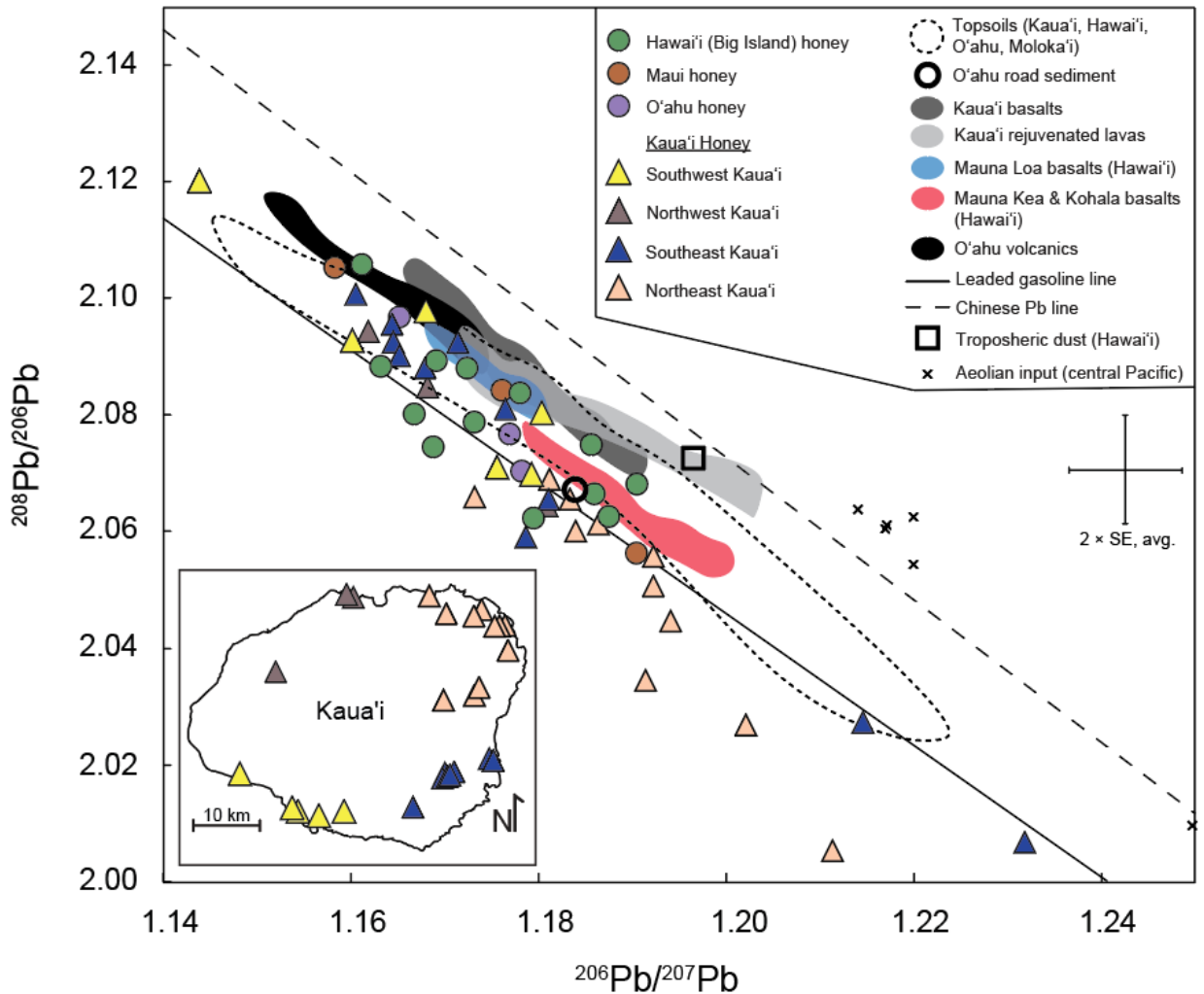


**Figure 5.** Panel A. Lead isotope plot of all honey analyzed in this study, with honey from previous studies included for context, as shaded fields (see Fig. 4 for symbol legend and caption for references). Data field labels denoted with an asterisk (\*) include bees (and other hive products) in addition to honey. The Sydney field includes data for bees, honey, and wax from European honeybees (*Apis mellifera*) and an Australian native bee species (*Tetragonula carbonaria*) (Zhou et al., 2018b, 2018a). The line (featured in all panels) is a mixing line between the Broken Hill and Southeast Missouri (SE MO) ore endmembers; a representation of the isotopic compositions of leaded gasolines used globally (Gulson, 1984; Sangster et al., 2000). The average error bar ( $2 \times$  standard error) is the mean error for all honey (except WI honey, for which error bars are smaller than the symbols). Panel B. Lead isotope plot of all honey data from this study (orange symbols in shaded yellow field), with various consumables from around



the globe for comparison: Australian wines (Kristensen et al., 2016), vinegars (apple cider, balsamic, wine, malt, cognac) (Ndung'u et al., 2011), and North American cigarettes (Guo et al., 2015). *Panel C.* Pb isotope plot of all honey data from this study (orange symbols in shaded yellow field), with the global aerosol array superimposed for comparison (Bollhöfer et al., 1999; Bollhöfer and Rosman, 2001, 2000).

The four galena samples from southwest Wisconsin have an isotopic range of 23.1996-23.8820, 16.1133-16.1797, and 42.750-43.281 for  $^{206}\text{Pb}/^{204}\text{Pb}$ ,  $^{207}\text{Pb}/^{204}\text{Pb}$ , and  $^{208}\text{Pb}/^{204}\text{Pb}$ , respectively (Table S2). These results are within the Pb isotopic range reported for other MVT Pb deposits from the Upper Mississippi Valley District and are in good agreement with other ores from southwest Wisconsin, which have notably radiogenic compositions (Fig. 7, 8) (Millen et al., 1995).



**Figure 6.** Lead isotope plot of honey collected from four Hawaiian Islands: Hawai'i, Maui, O'ahu,

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423 and Kaua'i. For context, Pb isotopic compositions of local geology is included: basalts from  
424 Mauna Loa, Mauna Kea, and Kohala volcanoes (Abouchami et al., 2000; Hanano et al., 2010;  
425 Weis et al., 2011), Kaua'i basalts and O'ahu volcanics (Wai'anae and Ko'olau) (Williamson et al.,  
426 2019). Topsoils (dotted field) are compiled from Monastra et al., (2004) and Spencer et al.,  
427 (1995) and O'ahu road sediments are from Sutherland et al. (2003). The gasoline mixing line is  
428 the same as described in Figure 5. The Chinese Pb line is defined by ores, coals, and fuel  
429 emissions from China (Bi et al., 2017; Cheng and Hu, 2010). Other aeolian inputs include  
430 samples collected from central Pacific Ocean sites surrounding the Hawaiian Islands (Jones et  
431 al., 2000) and tropospheric silicate dust collected on Hawai'i (Spencer et al., 1995). Inset map:  
432 Kaua'i honey sampling locations.  
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#### 434 **4. DISCUSSION**

##### 435 **4.1. Eastern North America and the Great Lakes region**

###### 436 *4.1.1. New York Metro Area*

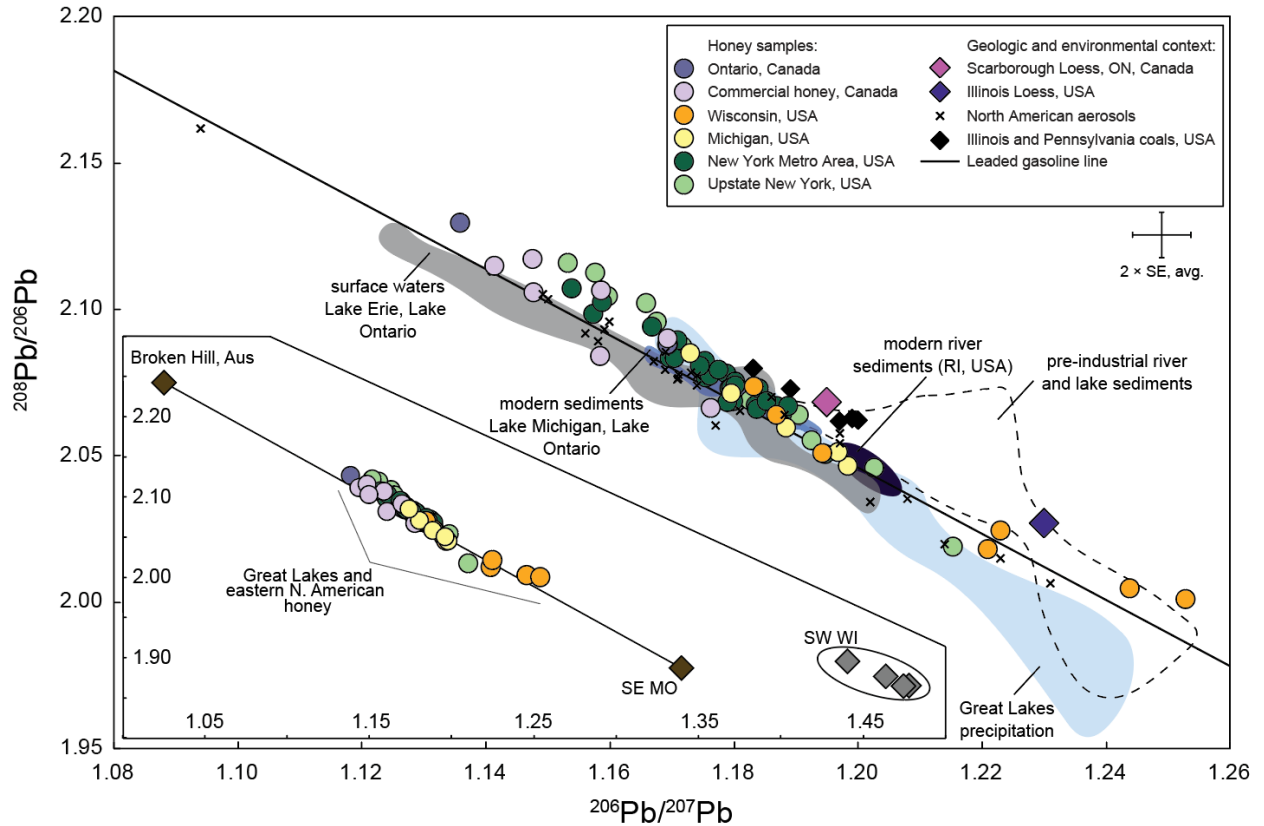
437 Cities tend to store more metals (especially Pb, as it lingers in the environment), since they host  
438 myriad industrial centers, roadways, ports, etc., in a concentrated area; the size and age of the  
439 city compound this effect (Chambers et al., 2016; Laidlaw and Filippelli, 2008; Mielke et al.,  
440 2011). Lead concentrations in the NY Metro honey exhibit a similar distribution of Pb in topsoil  
441 collected throughout NYC: higher concentrations in areas of high traffic, industrial activity, and  
442 in the oldest parts of the city, with the highest Pb concentrations measured in soils from  
443 Manhattan, Brooklyn, and Queens (Cheng et al., 2015; Li et al., 2018; Paltseva, 2019). The effect  
444 of nearby land use on other metals in honey in these regions is evident: for example, elevated  
445 Sb (vehicle brake wear) (Von Uexküll et al., 2005), Zn (galvanized structures, rubber tire crumbs)  
446 (Huber et al., 2015), and V (fuel oil combustion) (Zhao et al., 2013) (Fig. 2, S1). In a study of  
447 vegetables (root, leafy, and juicy varieties) from urban gardens across NYC, McBride et al.  
448 (2014) determined that Pb content is mostly a result of physical contamination of resuspended  
449 soil particulates and aerosol deposition, both of which are proven pathways for environmental  
450 particulates to make their way into honey (Negri et al., 2015; Pellicchia and Negri, 2018; Zhou  
451 et al., 2018b). Thus, typical city activities that contribute to emissions and particulate  
452 resuspension, such as heavy stop-and-go traffic, shipping ports, demolition, and construction,  
453 are all contributing to elevated levels of metals in NY Metro honey, relative to levels measured  
454 in honey collected further from the city center. Similar metal concentration gradients, as a  
455 function of distance from the city center, were observed in honey from Metro Vancouver,

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456 Canada, and Paris, France (Smith et al., 2020, 2019; Smith and Weis, 2020), with some logical  
457 differences (e.g., there is minimal spatial variation in V concentration in the non-port city of  
458 Paris, compared to New York and Vancouver that host major shipping ports) (Fig. 2, insets).

459 The Pb isotopic range of the NY Metro honey ( $^{206}\text{Pb}/^{207}\text{Pb}$ : 1.154-1.195) is consistent  
460 with ranges previously reported for topsoil, dust, and paints throughout the NY Metro Area.  
461 Some honey from this study is slightly (not significantly) less radiogenic than that of outdoor  
462 dust collected throughout the city from the tops of pedestrian crossing signals at major  
463 intersections ( $^{206}\text{Pb}/^{207}\text{Pb}$ : 1.17-1.23) (Caravanos et al., 2006). Our honey range overlaps with  
464 that of house paints ( $^{206}\text{Pb}/^{207}\text{Pb}$ : 1.141-1.186) and indoor dust samples ( $^{206}\text{Pb}/^{207}\text{Pb}$ : 1.166-  
465 1.192) collected in Hillsdale, NJ (part of the NY Metro Area) (Jaeger et al., 1998). The NY Metro  
466 honey data also falls within the Pb isotopic ranges measured in indoor house dust ( $^{206}\text{Pb}/^{207}\text{Pb}$ :  
467 1.134-1.227), interior paints (1.092-1.267), garden soils (1.612-1.200), and street dust (1.155-  
468 1.199) collected in Jersey City, NJ (immediately west of Lower Manhattan, across the Hudson  
469 River, Adgate et al., 1998), and topsoil from Staten Island ( $^{206}\text{Pb}/^{207}\text{Pb}$ : 1.144-1.196, Pribil et al.,  
470 2014).

471 While the Pb isotopic compositions of NY Metro honey fall within the expected isotopic  
472 range of the region (Fig. 7), the results do not reveal a predictable geospatial gradient or trend  
473 at city-scale, in contrast to previous studies in Metro Vancouver (Smith et al., 2019; Smith and  
474 Weis, 2020) and Sydney, Australia (Zhou et al., 2018b). This is likely due to NYC being much  
475 larger and older than both Metro Vancouver and Sydney, meaning modern and historical land  
476 use, zoning, and industry are comparatively more extensive and convoluted in NYC. Indeed, a  
477 similar honey survey, recently undertaken in an even older city (Paris, France), showed a  
478 completely predictable Pb concentration gradient (matching the footprint of the fallout from  
479 the 2019 fire at Notre-Dame cathedral), while the Pb isotopic compositions of honey exhibited  
480 no geospatial trends within the city due to millennia-scale history of Pb use in Paris (Smith et  
481 al., 2020). The Paris study and the NY metro honey results imply that the utility of km-scale  
482 geospatial resolution of Pb isotopic compositions in honey is not only a function of land use and  
483 recent historical development but of the comprehensive geochemical ‘baseline’ of the region,  
484 i.e., all natural and anthropogenic Pb inputs, to date (Matschullat et al., 2000).



**Figure 7.** Honey from the Great Lakes region/eastern North America. For commercial honeys from Canada, we only included honey brands that claimed to be of Canadian origin (more specific origin information is unknown). For other honey, if specific origin location is known, it is listed in Table S1. Main plot: Selected local and regional context is provided from the literature, including North American aerosols (Bollhöfer and Rosman, 2001), loess (Biscaye et al., 1997), coals from Illinois and Pennsylvania (Chow and Earl, 1972), Great Lakes region precipitation (Sherman et al., 2015), Great Lakes (Erie and Ontario) surface waters (Flegal et al., 1989), modern (1990s and later) lake sediments from inland lakes in Michigan and Ontario (Cheyne et al., 2018), and modern sediments (1980s and later) from the Pettaquamscutt River (Lima et al., 2005). The pre-industrial field (dashed line) consists of lake and river sediments from the 1700s (Cheyne et al., 2018; Graney et al., 1995; Lima et al., 2005). The gasoline mixing line is the same as described in Figure 5. Average error bars ( $2 \times$  standard error) are for all honey (except the WI honey, which has error bars smaller than the symbols). Inset: A simplified version of the main panel, zoomed out to show endmembers of the gasoline mixing line and other Pb ores. 'SW WI' is galena from a MVT ore deposit in southwest Wisconsin (Lafayette Co.).

#### 4.1.2. Pb isotopes in USA versus Canadian honey

Lead isotopic compositions of honey from eastern USA and Canada are comparable to Pb isotopic ranges reported in other environmental matrices in that region: precipitation, snow, and modern lake sediments and surface waters, and all honey from this region plots within

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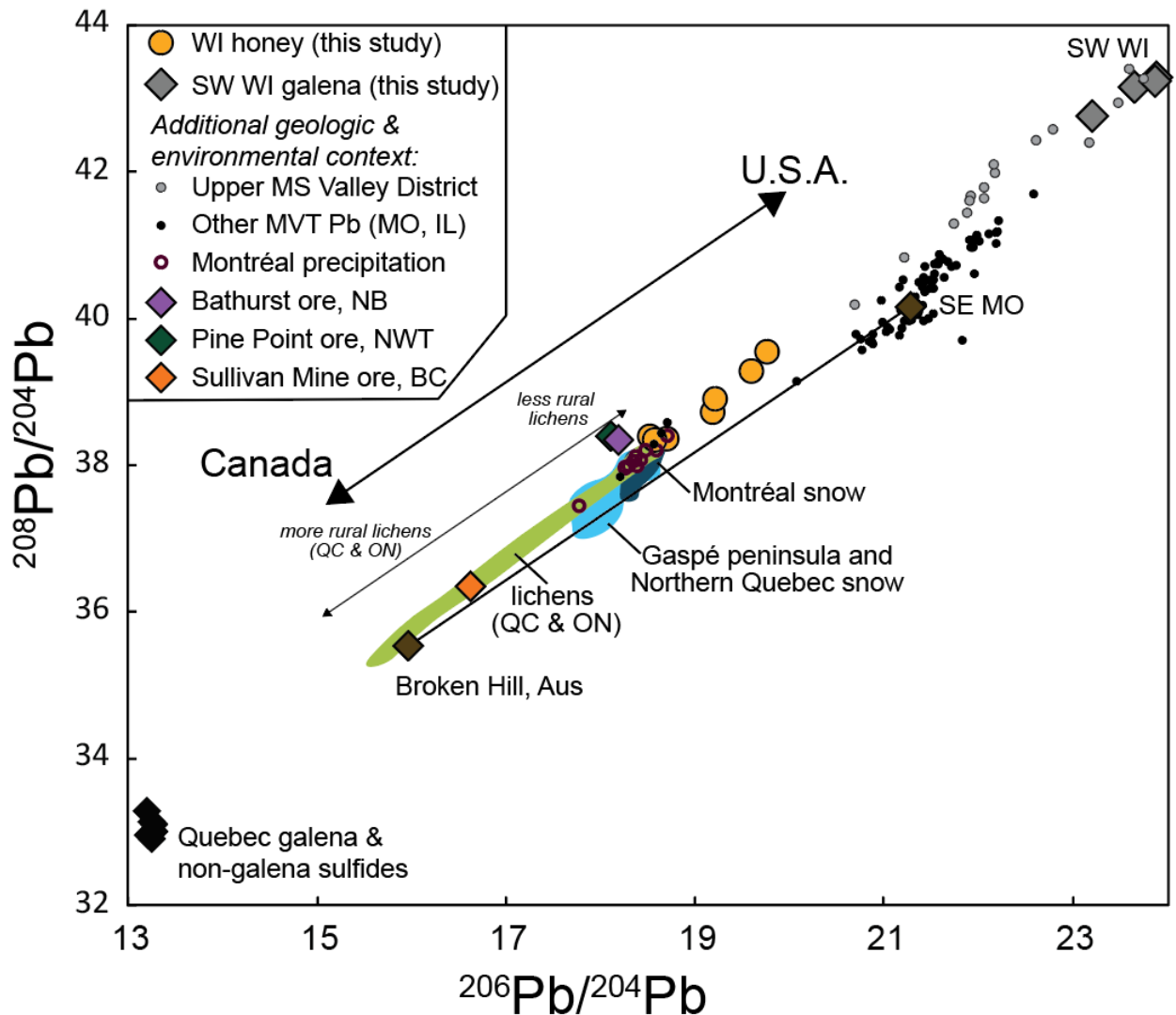
506 error of the leaded gasoline mixing line (Fig. 7 and references therein). Honey samples from the  
507 USA have Pb isotopic compositions that are more radiogenic than those from Canada ( $p_{adj.} <$   
508  $0.001$  for  $^{206}\text{Pb}/^{207}\text{Pb}$  and  $^{208}\text{Pb}/^{206}\text{Pb}$ ). This can be explained by differing industrial histories in  
509 these regions and different sources of raw materials for Pb smelting, manufacturing, and  
510 recycling in each country. For instance, both countries used some amount of leaded gasoline  
511 mixed with tetraethyl Pb (TEL) produced from Southeast Missouri (SE MO) Pb ore, but Canada  
512 used only a small percentage in comparison to the USA and sourced the majority of their TEL  
513 from British Columbia and New Brunswick (Canada imported  $< 1\%$  of Pb ore used for TEL),  
514 while the USA supplemented their SE MO Pb with ores from Australia, Mexico, and Peru  
515 (Sturges and Barrie, 1987). The USA also had considerably more industrial activity, consuming  
516 more Pb and other raw materials than Canada (from  $\sim 1750$  to the present), resulting in  
517 present-day Pb enrichment factors ( $EF_{\text{Pb}}$ ) in lake sediments of  $\sim 100$  in the Great Lakes region  
518 (USA side) compared to  $EF_{\text{Pb}}$  of  $\sim 10$  in lake sediments from eastern Canada (Marx et al., 2016).

519 In the 1980s, when leaded gasoline was still in widespread use, dissolved Pb samples  
520 from Lakes Erie and Ontario had more radiogenic compositions when the wind was blowing  
521 from the south/USA and were conversely less radiogenic when the wind was blowing from the  
522 north/Canada (Flegal et al., 1989). Today, it seems that ongoing sulfide mining and Pb refining  
523 in the Mississippi Valley and Quebec (Abitibi Greenstone Belt) continue to ensure that honey  
524 from the eastern USA and Canada have differing Pb isotopic compositions, especially since the  
525 major metropolitan areas of the USA are downwind of key industrial and mining regions of the  
526 mid- and eastern USA, due to prevailing winds (Klink, 1999). This is further clarified when using  
527 the  $^{204}\text{Pb}$  isotope to examine this disparity; honey from WI plots closer to the USA/MVT Pb ores,  
528 as expected based on context from USA and Canadian Pb ores and environmental samples  
529 (precipitation, snow, lichens) from both rural and urban areas in eastern Canada (Fig. 8 and  
530 references therein).

531 Earlier studies that focused on city-scale nuance of Pb distribution in honey  
532 demonstrated that honey can resolve input from point sources or acute Pb pollution events  
533 (Smith et al., 2019; Smith and Weis, 2020; Zhou et al., 2018b), but this localized variation is lost  
534 at the regional/sub-continental scale, which better reflects regional-scale processes, i.e.,

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535 different Pb ore endmembers used in the eastern USA versus eastern Canada in this case. Metal  
536 concentrations and Pb isotopic compositions of sediment cores from southern Ontario and  
537 northern Michigan recorded the history of local and regional metal input, which, at certain  
538 points in time, overwhelmed that of the continental, global leaded gasoline trends of the  
539 twentieth century (Cheyne et al., 2018). Presumably, if honey had been collected at a higher  
540 resolution near Great Lakes industrial centers throughout the last century (e.g., Toronto,  
541 Chicago, Milwaukee, Toledo), it would reveal the local nuances of industrial pollution at each  
542 site.



543  
544 **Figure 8.** Lead isotope plot featuring honey from south-central WI. The honey plots toward the  
545 USA end of the Pb mixing line defined by differing emissions in eastern Canada vs. eastern USA,  
546 e.g., after Simonetti et al. (2000a). Reference reservoirs include SW WI galena (this study),  
547 Upper Mississippi (MS) Valley District MVT ores (Millen et al., 1995), other USA MVT Pb ores

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(Goldhaber et al., 1995), Canadian (Quebec) Pb ores (galena and non-galena sulfides) from the Atibiti Pb sulfide belt (Franklin et al., 1983), and other economically significant Pb ores (mined throughout the past century) in Canada: Pb from the Sullivan Mine (Kimberley, BC), the Pine Point Mine (Great Slave Lake, NWT), and the Bathurst Mining District (northeast NB) (compiled by Sangster et al., 2000). Environmental context includes Montréal precipitation data (Simonetti et al., 2000b), and snowpack and lichen data from Montréal, moderately remote areas (e.g., the Gaspé Peninsula/St. Lawrence Valley), and very remote areas (i.e., along a north-south transect between southern ON/QC and Hudson Bay) (Carignan and Gariépy, 1995; Simonetti et al., 2000a). Error bars ( $2 \times$  standard error) for the WI honey and southwest WI galena data are smaller than the symbols.

#### 4.2 Kaua’i, Hawaiian Islands

Major sources of Pb on the Hawaiian Islands are lithogenic (Hawaiian basalts), and legacy anthropogenic (leaded gasoline, leaded paint, discontinued/banned pesticides, e.g., lead arsenate). There is also some aeolian dust input from northeast Asia, confirmed by its continental-derived  $^{87}\text{Sr}/^{86}\text{Sr}$  ratios (too high to be derived from the basaltic, Hawaiian geology) measured in Hawaiian topsoil (Kurtz et al., 2001; Spencer et al., 1995). Lead source apportionment calculations for road-deposited sediments from O’ahu indicate inputs from gasoline additives of 31-81 % in sediment collected after 1968, the period when anthropogenic Pb input was more easily identified, since TEL content of USA gasoline shifted toward increased usage of the very radiogenic SE MO ores (Sutherland et al., 2003). Compared to honey from the other Hawaiian Islands, Kaua’i honey exhibits the largest range in Pb isotopes and the most deviation from lithogenic Pb isotopic compositions, including basalts from Kaua’i and other Hawaiian islands (Hanano et al., 2010; Williamson et al., 2019) and loess transported from the Chinese Loess Plateau and deposited in the central Pacific (Jones et al., 2000). The largest Pb isotopic ranges are observed in honey from the eastern side of the island, where most of Kaua’i’s small population (~72,200 people total) reside, so the observed isotopic compositions are likely a result of current and legacy anthropogenic activity (e.g., backyard scrapping of old vehicles, old paint, legacy gasoline, agricultural activity), especially since any deviations from the global gasoline mixing line (i.e., honey from northeast Kaua’i) do not plot in the direction of Kaua’i’s natural/geogenic Pb isotopic compositions. For comparison, commercially available honey from Hawai’i Is. has Pb isotopic compositions that plot between natural and anthropogenic Pb isotopic compositions (Fig. 6).

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581 Large-scale agricultural activity is likely contributing to the anomalously high metal  
582 concentrations observed in Kaua'i honey compared to honey from O'ahu, Maui, and Hawai'i  
583 (Fig. 3, S2). The south side of Kaua'i is mostly in seed crop and coffee production (Spengler et  
584 al., 2019). A previous study of Kaua'i honey established that concentrations of glyphosate (an  
585 organophosphate herbicide marketed as Roundup™ by Monsanto/Bayer AG) measured in  
586 honey correlated well with land use: hives near large-scale agriculture operations, roadsides, or  
587 golf courses produced honey with higher concentrations of glyphosate (Berg et al., 2018). The  
588 regions where Berg et al. (2018) observed the highest glyphosate concentrations in honey are  
589 in the south/southwest of the island, where we observe some of the highest metal  
590 concentrations (in particular, Pb, Ti, V, Cr, Fe, Ga, Zr, Ba, Fig. 3, S2) observed in honey, not only  
591 from the Hawaiian Islands, but in any honey analyzed for this study (Figs. 1, 4, Table S1).  
592 Pesticide and fertilizer use as a source for some of these metals should not be discounted,  
593 including Cd, Co, Cu, Ni, Pb, and Zn (Gimeno-García et al., 1996). Defarge et al. (2018) measured  
594 As, Co, Cr, Ni, and Pb (ranging from 10s to 100s ng/g) as formulant ingredients in glyphosate-  
595 based herbicides and other commercially available herbicides, fungicides, and insecticides.  
596 Fertilizer, including phosphate fertilizers, sewage sludges, manures, and fly ash (compiled by  
597 Kabata-Pendias, 2010 and U.S. EPA, 1999), also introduces metals of foreign origin into  
598 agricultural fields. Additionally, three honey samples collected in southern Kaua'i were from  
599 feral hives, one within an abandoned building, all three within 0.5 km of abandoned farming  
600 vehicles and equipment.

601 Another important contributing factor for metal transport is climate. Kaua'i's central,  
602 high-elevation regions experience large amounts of rainfall; up to 1140 cm per year on Mt.  
603 Waialeale, among the highest amounts measured anywhere on Earth (US National Weather  
604 Service, 2020), which leads to immense amounts of erosion and washout into the valleys  
605 (Ferrier et al., 2013) even on the comparatively 'dry' south/southwest side of the island. The  
606 Waimea River, which empties into Waimea Bay (mean daily discharge rate of 1.1 m<sup>3</sup>/s, 63-years  
607 of data, USGS, 2020), is fed by a high-elevation swamp that receives significant amounts of  
608 rainfall. Coupled with a humid, tropical climate, extreme precipitation eventually reduces soil  
609 pH and mobilizes heavy metals from soils (Spencer et al., 1995). On O'ahu, others (Andrews and



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610 Sutherland, 2004; Heinen De Carlo and Anthony, 2002) observed that metals of both  
611 anthropogenic and lithogenic origin were transported downstream where they gradually  
612 accumulated in the upper soil horizons and sediment layers relative to background levels, so a  
613 similar transport mechanism is likely on Kaua'i (which has undergone a similar geologic and soil-  
614 forming history) (e.g., Orazio et al., 2007).

615 A new, modern study on the geochemistry of Kaua'i topsoil, C-horizon soils, sediments  
616 from Waimea Bay (downstream of the Waimea Valley), e.g., after Spencer et al. (1995), and  
617 honey would be worthwhile. In particular, sediments would offer the temporal context  
618 necessary to reveal whether the downstream sediments contained lower metal contents prior  
619 to large-scale farming efforts in the past decades. The ongoing effects of metal and pesticide  
620 transport in this region and assessment of the polluting role (if any) that other unique land use  
621 features on Kaua'i (e.g., the US Navy Pacific Missile Range Facility, the Port Allen solar voltaic  
622 power plant, and legacy diesel-fed sugar mills) may have on the region should remain the  
623 subject of ongoing and future studies. Higher-resolution honey sampling on the other Hawaiian  
624 Islands would provide a useful comparison between the Pb isotopic compositions and metal  
625 concentrations observed on Kaua'i versus those in larger Hawaiian urban centers (e.g.  
626 Honolulu, O'ahu).

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### 628 **4.3 Honey and the global Pb array**

629 The Pb isotopic compositions of honey fit predictably on the global Pb isotope array based on  
630 the geographic origin of the honey. For example, honey samples from New Zealand have Pb  
631 isotopic compositions that are distinctly less radiogenic compared to honey from western  
632 Europe, Canada, Hawai'i, and the continental USA, while New Zealand compositions are  
633 comparable to other samples from Oceania: bees, honey, beeswax, topsoil, and dust from  
634 Sydney, Australia, red wines from South Australia (Kristensen et al., 2016), and aerosols from  
635 New Zealand, Tasmania, and Australia (Bollhöfer and Rosman, 2000). The least radiogenic Pb  
636 isotopic compositions reported for honey, bees, and other hive products were collected in the  
637 vicinity of the Broken Hill Mine, Australia (Zhou et al., 2018b), an area noted for providing one  
638 of the two Pb ores used to produce the majority of the world's TEL (the other being SE MO ore,

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4 639 at the other end of the Pb isotopic composition spectrum for TEL used globally) (Fig. 5a).

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6 640 Western European honey from this study (Belgium and Germany) have similar Pb  
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8 641 isotopic compositions to previously reported values for honey from France (from Paris and the  
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10 642 Auvergne-Rhône-Alpes regions, Smith et al. 2020) and have similar Pb isotopic compositions to  
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12 643 wines and vinegars from western Europe (France, Spain, Italy) (Fig. 5b, and references therein).  
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14 644 Honey and other products from the USA tend to have the most radiogenic Pb isotopic  
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16 645 compositions, with honey from Canada falling between the USA and western European ranges.  
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18 646 This is likely due to consumption of TEL from differing sources throughout the twentieth  
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20 647 century. There are few continental-scale isoscapes available for Pb, but the honey results match  
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22 648 isoscapes mapped using other environmental proxies: the Pb isotopic compositions for topsoil  
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24 649 and tooth enamel from the USA are generally more radiogenic than those observed in western  
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26 650 European isoscapes constructed using similar proxies (Keller et al., 2016; Reimann et al., 2012,  
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28 651 2011).

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30 652 Unsurprisingly, the limited honey samples from regions other than North America,  
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32 653 Western Europe, and New Zealand (Table S1, Fig. 5a) also plot along the global Pb array. These  
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34 654 are the first Pb isotopic compositions reported for honey from Afghanistan, Brazil, Cuba,  
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36 655 Germany, Liberia, Taiwan, and Turkey and may serve as a starting point for future studies in  
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38 656 those regions. Confirming the geographic authenticity of these honey samples is beyond the  
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40 657 scope of this study, but their Pb isotopic compositions fit well with existing Pb aerosol data  
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42 658 from the same region/country (Fig. 5). For example, the two Cuban honey blends and three  
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44 659 Brazilian honey samples analyzed in this study have Pb isotopic compositions in agreement with  
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46 660 ranges reported for aerosols in Cuba and Brazil (collected at the end of the twentieth century)  
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48 661 (Bollhöfer and Rosman, 2000).

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#### 50 51 663 **4.4 Geochemical implications of sampling resolution**

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53 664 A city-scale survey of metals in honey provides insight into factors that influence metal  
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55 665 distribution locally, and the corresponding Pb isotopic compositions can provide insight into  
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57 666 metal sources. These factors may include local topography, land use, municipal-level  
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59 667 environmental policy and regulation, construction, and acute pollution events (Smith and Weis,

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668 2020). For example, the metal concentration data for honey from the New York Metro Area and  
669 Kaua'i reveal interesting local patterns of metal distribution. The metal concentrations reported  
670 in this study were selected for their utility in geochemical investigations; these metals are most  
671 useful on a local scale when combined with a high sampling resolution to differentiate between  
672 anthropogenic input versus local background, i.e., best for sourcing metals as opposed to  
673 tracing the geologic source of the honey. Thus, these metals are not directly useful for  
674 comparison between global regions since there is too much variation in metal mobility as a  
675 result of local factors such as plant diversity and soil properties (pH, porosity, organic material,  
676 and clay content) (Kabata-Pendias, 2010, 2004). The suite of metals reported in this study may  
677 nonetheless supplement future honey surveys that include additional analytes that do assist in  
678 resolving the region of origin for honey or to confirm honey authenticity and/or botanical  
679 origin. These include certain macronutrients (K, P) (Chen et al., 2014; Zhou et al., 2018c) and  
680 lanthanide series metals (Pellerano et al., 2012).

681

#### 682 *4.4.1 Geochemical 'baseline' and future of Pb isotope use in honey*

683 Honey meets the requirements suggested by Darnley et al. (1995) for adding viable data to  
684 a global geochemical database. These requirements include being a commonly available  
685 sampling matrix where adequate amounts can be stored indefinitely for future reference. Lead  
686 isotopic compositions of wine samples from France (Epova et al., 2020) and Australia  
687 (Kristensen et al., 2016) with vintages spanning the twentieth-century exhibit a temporal trend,  
688 reflecting the introduction and subsequent phase-out of leaded gasoline in their respective  
689 geographical regions. Presumably, if honey archives existed throughout the past two centuries,  
690 they would also reflect the key moments of Pb use throughout recent history. Like wine, honey  
691 is easily stored for very long periods of time, if not indefinitely. Given that the most recent and  
692 advanced spectroscopic techniques (e.g., NMR, FT-IR) are becoming common-place for rapid-  
693 throughput authentication efforts (Wu et al., 2017), perhaps archiving of honey will become  
694 standard practice, creating a chemical repository for various environmental investigations, i.e.,  
695 as a record of past environmental conditions and a continuous record of baseline compositions.

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696 Establishing a baseline (natural/background input + anthropogenic input to-date) for a  
697 given chemical species is important in geochemical surveys (Matschullat et al., 2000; Smith et  
698 al., 2018). Since these data reflect the current baseline for Pb isotopic compositions in honey,  
699 they represent a useful tool for future comparison as Pb use shifts. However, confidently  
700 identifying the contributions of natural *versus* anthropogenic Pb in consumer products in the  
701 modern, global economic climate is becoming increasingly difficult for three key reasons. First,  
702 recent (since the mid-1800s) anthropogenic input of Pb into the environment is large (1 to 2  
703 orders of magnitude larger) compared to natural inputs (Han et al., 2002; Nriagu, 1989; Weiss  
704 et al., 1999). Second, all anthropogenic sources were originally derived from natural sources –  
705 this is obvious but depending on the scale of a given Pb investigation (local/city-scale,  
706 regional/continental, global), this concept can conflate Pb mixing models unless possible Pb  
707 inputs are identified and quantified at each scale (Flegal and Odigie, 2020). Third, ores and raw  
708 materials are shipped globally for refining and manufacturing, then manufactured goods are  
709 shipped elsewhere, and post-consumer materials (e.g., Pb batteries) may be shipped  
710 internationally yet again for recycling. It is important to note that even raw materials from the  
711 same ore deposit can exhibit natural Pb isotopic variability (e.g., Sangster et al., 2000). The food  
712 industry has also been globalized, in particular for specialty, non-perishable items like honey  
713 where claims of ‘terroir’ or source environment can be economically valuable, e.g., the manuka  
714 honey industry of New Zealand (Carter et al., 2016; Rogers et al., 2014). In general, the honey  
715 market remains non-transparent. For example, the distributor of commercially-available honey  
716 purchased in North Carolina, USA, for this study stated that the product was a blend of  
717 American honey and imported honey but would not disclose any further details about mixing  
718 ratios or the geographic source locations of the honey (Pb isotopic composition of this honey is  
719 consistent with other North American honey, however).

720 For honey and other biomonitoring applications, we suggest using Pb isotopes in  
721 conjunction with other tools to strengthen sourcing studies. For example, analysis of Sr or Nd  
722 isotopes in honey may help distinguish geogenic inputs, e.g., dust and endmember loesses  
723 (Grousset and Biscaye, 2005), from anthropogenic inputs (Baroni et al., 2015; Voerkelius et al.,  
724 2010). We also suggest more extensive use of the <sup>204</sup>Pb isotope in environmental Pb surveys

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4 725 (e.g., Fig. 8) to verify that Pb sourcing interpretations are not oversimplified (e.g., Ellam, 2010).  
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6 726 If the goal is to trace the origin of the honey, rather than the source of the Pb and other metals,  
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8 727 there is no doubt that a global geochemical database for honey (in which sampling origin is  
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10 728 known and knowledge exists regarding anthropogenic and geogenic sources of metals in honey)  
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12 729 will supplement and strengthen the existing statistical approaches that use carbon isotopes and  
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14 730 trace metal data (Zhou et al., 2018c) and the higher-throughput spectroscopy methods of  
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16 731 honey sourcing and authentication, e.g., various infrared and nuclear magnetic resonance  
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18 732 techniques (Wu et al., 2017).  
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## 21 734 **5. CONCLUSIONS**

23 735 This study presents Pb isotopic and metal concentration data for honey at three different  
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25 736 scales: local, with high-resolution sampling (NY Metro Area and Kaua'i), regional (Eastern North  
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27 737 America), and global. Local-scale honey data gives insight into local processes affecting metal  
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29 738 distribution and aids in identifying point and/or diffuse sources of metals. Metal concentrations  
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31 739 in honey indicate potential effects of large-scale farming in and around Waimea on Kaua'i and  
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33 740 typical, city-type distribution gradients of metals associated with human activity (e.g., Pb, Sb, V,  
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35 741 Zn) throughout New York City. On these small geographic scales, there is some variation in Pb  
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37 742 isotopic composition, but their range is largely contained within the isotopic constraints defined  
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39 743 by local sources of Pb, whether natural or anthropogenic. In larger-scale geochemical surveys  
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41 744 using honey, metal concentrations are less useful for mapping distributions and identifying  
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43 745 point sources and the Pb isotopic compositions reflect regional and global processes related to  
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45 746 Pb. A regional survey of Pb isotopic compositions in honey reflects the industry, urbanization,  
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47 747 and geology of eastern North America, revealing differences in modern Pb emissions in the  
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49 748 eastern USA and eastern Canada. Globally, Pb isotopes in honey roughly reflect the continental  
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51 749 origin of the honey: samples from the USA are most radiogenic and honey from New Zealand is  
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53 750 least radiogenic. The global trend predictably matches the modern global Pb array, as defined  
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55 751 by aerosols, varying regional geologic background, and nearly a century of leaded gasoline use  
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57 752 around the world. This study provides a sound starting point for building current baseline data  
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59 753 for honey as a biomonitor for metals in the regions discussed here and provides a global

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4 754 framework for Pb isotopic compositions in honey, which is useful context for more localized  
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6 755 studies. This study also demonstrates the utility of a community science effort for honey  
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8 756 collection.  
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#### 10 757 **SUPPLEMENTARY MATERIAL**

11 758 -Tables S1-S3

12 759 -Figures S1-S2

#### 13 760 14 761 15 762 **ACKNOWLEDGEMENTS**

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#### 29 776 **AUTHOR CONTRIBUTIONS**

30 777 **K. Smith:** investigation, formal analysis, validation, writing-original draft, writing-review and  
31 778 editing, visualization, resources **D. Weis:** conceptualization, funding acquisition, writing-review  
32 779 and editing, supervision **S. Scott:** resources, investigation, validation, writing-review and editing  
33 780 **C. Berg, Y. Segal, P. Claeys:** resources, writing-review and editing.  
34 781

#### 35 782 **CONFLICT OF INTEREST**

36 783 The authors declare no conflicts of interest relevant to this study.  
37 784

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#### 45 792 **REFERENCES**

46 793 Abouchami, W., Galer, S.J.G., Hofmann, A.W., 2000. High precision lead isotope systematics of  
47 794 lavas from the Hawaiian Scientific Drilling Project. *Chem. Geol.* 169, 187–209.  
48 795 [https://doi.org/10.1016/S0009-2541\(00\)00328-4](https://doi.org/10.1016/S0009-2541(00)00328-4)  
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796 Adgate, J.L., Rhoads, G.G., Liroy, P.J., 1998. The use of isotope ratios to apportion sources of lead  
797 in Jersey City, NJ, house dust wipe samples. *Sci. Total Environ.* 221, 171–180.  
798 [https://doi.org/10.1016/S0048-9697\(98\)00282-4](https://doi.org/10.1016/S0048-9697(98)00282-4)

799 Andrews, S., Sutherland, R.A., 2004. Cu, Pb and Zn contamination in Nuuanu watershed, Oahu,  
800 Hawaii. *Sci. Total Environ.* 324, 173–182. <https://doi.org/10.1016/j.scitotenv.2003.10.032>

801 Annual Estimates of the Resident Population: April 1, 2010 to July 1, 2016 - Combined Statistical  
802 Area; and for Puerto Rico -2016 Population Estimates, U.S. Census Bureau.

803 Baroni, M. V., Podio, N.S., Badini, R.G., Inga, M., Ostera, H.A., Cagnoni, M., Gautier, E.A., García,  
804 P.P., Hoogewerff, J., Wunderlin, D.A., 2015. Linking soil, water, and honey composition to  
805 assess the geographical origin of Argentinean honey by multielemental and isotopic  
806 analyses. *J. Agric. Food Chem.* 63, 4638–4645. <https://doi.org/10.1021/jf5060112>

807 Batista, B.L., da Silva, L.R.S., Rocha, B.A., Rodrigues, J.L., Berretta-Silva, A.A., Bonates, T.O.,  
808 Gomes, V.S.D., Barbosa, R.M., Barbosa, F., 2012. Multi-element determination in Brazilian  
809 honey samples by inductively coupled plasma mass spectrometry and estimation of  
810 geographic origin with data mining techniques. *Food Res. Int.* 49, 209–215.  
811 <https://doi.org/10.1016/j.foodres.2012.07.015>

812 Berg, C.J., Peter King, H., Delenstarr, G., Kumar, R., Rubio, F., Glaze, T., 2018. Glyphosate residue  
813 concentrations in honey attributed through geospatial analysis to proximity of large-scale  
814 agriculture and transfer off-site by bees. *PLoS One* 13, 1–18.  
815 <https://doi.org/10.1371/journal.pone.0198876>

816 Bi, X.Y., Li, Z.G., Wang, S.X., Zhang, L., Xu, R., Liu, J.L., Yang, H.M., Guo, M.Z., 2017. Lead Isotopic  
817 Compositions of Selected Coals, Pb/Zn Ores and Fuels in China and the Application for  
818 Source Tracing. *Environ. Sci. Technol.* 51, 13502–13508.  
819 <https://doi.org/10.1021/acs.est.7b04119>

820 Biscaye, P.E., Grousset, F.E., Revel, M., Van Der Gaast, S., Zielinski, G.A., Vaars, A., Kukla, G.,  
821 1997. Asian provenance of glacial dust (stage 2) in the Greenland Ice Sheet Project 2 Ice  
822 Core, Summit, Greenland. *J. Geophys. Res. Ocean.* 102, 26765–26781.  
823 <https://doi.org/10.1029/97JC01249>

824 Bollhöfer, A., Chisholm, W., Rosman, K.J.R., 1999. Sampling aerosols for lead isotopes on a  
825 global scale. *Anal. Chim. Acta* 390, 227–235. [https://doi.org/10.1016/S0003-2670\(99\)00182-8](https://doi.org/10.1016/S0003-2670(99)00182-8)

826

827 Bollhöfer, A., Rosman, K.J.R., 2001. Isotopic source signatures for atmospheric lead: The  
828 Northern Hemisphere. *Geochim. Cosmochim. Acta* 65, 1727–1740.

829 Bollhöfer, A., Rosman, K.J.R., 2000. Isotopic source signatures for atmospheric lead: The  
830 Southern Hemisphere. *Geochim. Cosmochim. Acta* 64, 3251–3262.  
831 [https://doi.org/10.1016/S0016-7037\(00\)00436-1](https://doi.org/10.1016/S0016-7037(00)00436-1)

832 Bromenshenk, J.J., Carlson, S.R., Simpson, J.C., Thomas, J.M., 1985. Pollution monitoring of  
833 Puget Sound with Honey Bees. *Science.* 227, 632–634.

834 Caravanos, J., Weiss, A.L., Blaise, M.J., Jaeger, R.J., 2006. A survey of spatially distributed  
835 exterior dust lead loadings in New York City. *Environ. Res.* 100, 165–172.  
836 <https://doi.org/10.1016/j.envres.2005.05.001>

837 Carignan, J., Gariépy, C., 1995. Isotopic composition of epiphytic lichens as a tracer of the  
838 sources of atmospheric lead emissions in southern Québec, Canada. *Geochim. Cosmochim.*  
839 *Acta* 59, 4427–4433. [https://doi.org/10.1016/0016-7037\(95\)00302-G](https://doi.org/10.1016/0016-7037(95)00302-G)

1  
2  
3  
4 840 Carter, D.A., Blair, S.E., Cokcetin, N.N., Bouzo, D., Brooks, P., Schothauer, R., Harry, E.J., 2016.  
5 841 Therapeutic manuka honey: No longer so alternative. *Front. Microbiol.* 7, 1–11.  
6 842 <https://doi.org/10.3389/fmicb.2016.00569>  
7  
8 843 Chambers, L.G., Chin, Y.-P., Filippelli, G.M., Gardner, C.B., Herndon, E.M., Long, D.T., Lyons,  
9 844 W.B., Macpherson, G.L., McElmurry, S.P., McLean, C.E., Moore, J., Moyer, R.P., Neumann,  
10 845 K., Nezat, C.A., Soderberg, K., Teutsch, N., Widom, E., 2016. Developing the scientific  
11 846 framework for urban geochemistry. *Appl. Geochemistry* 67, 1–20.  
12 847 <https://doi.org/10.1016/j.apgeochem.2016.01.005>  
13  
14 848 Chen, H., Fan, C., Chang, Q., Pang, G., Hu, X., Lu, M., Wang, W., 2014. Chemometric  
15 849 determination of the botanical origin for Chinese honeys on the basis of mineral elements  
16 850 determined by ICP-MS. *J. Agric. Food Chem.* 62, 2443–2448.  
17 851 <https://doi.org/10.1021/jf405045q>  
18  
19 852 Cheng, H., Hu, Y., 2010. Lead (Pb) isotopic fingerprinting and its applications in lead pollution  
20 853 studies in China: A review. *Environ. Pollut.* 158, 1134–1146.  
21 854 <https://doi.org/10.1016/j.envpol.2009.12.028>  
22  
23 855 Cheng, Z., Paltseva, A., Li, I., Morin, T., Huot, H., Egendorf, S., Su, Z., Yolanda, R., Singh, K., Lee,  
24 856 L., Grinshtein, M., Liu, Y., Green, K., Wai, W., Wazed, B., Shaw, R., 2015. Trace metal  
25 857 contamination in New York City garden soils. *Soil Sci.* 180, 167–174.  
26 858 <https://doi.org/10.1097/SS.000000000000126>  
27  
28 859 Cheyne, C.A.L., Thibodeau, A.M., Slater, G.F., Bergquist, B.A., 2018. Lead isotopes as particulate  
29 860 contaminant tracers and chronostratigraphic markers in lake sediments in northeastern  
30 861 North America. *Chem. Geol.* 477, 47–57. <https://doi.org/10.1016/j.chemgeo.2017.11.043>  
31  
32 862 Chow, T.J., Earl, J.L., 1972. Lead Isotopes in North American Coals. *Science.* 176, 510–511.  
33 863 <https://doi.org/10.1017/CBO9781107415324.004>  
34  
35 864 Codling, G., Al Naggar, Y., Giesy, J.P., Robertson, A.J., 2016. Concentrations of neonicotinoid  
36 865 insecticides in honey, pollen and honey bees (*Apis mellifera* L.) in central Saskatchewan,  
37 866 Canada. *Chemosphere* 144, 2321–2328.  
38 867 <https://doi.org/10.1016/j.chemosphere.2015.10.135>  
39  
40 868 Conti, M.E., Botrè, F., 2001. Honeybees and their products as potential bioindicators of heavy  
41 869 metals contamination. *Environ. Monit. Assess.* 69, 267–282.  
42  
43 870 Darnley, A.G., 1995. International geochemical mapping - a review. *J. Geochemical Explor.* 55,  
44 871 5–10.  
45  
46 872 de Oliveira, R.C., Queiroz, S.C. do N., da Luz, C.F.P., Porto, R.S., Rath, S., 2016. Bee pollen as a  
47 873 bioindicator of environmental pesticide contamination. *Chemosphere* 163, 525–534.  
48 874 <https://doi.org/10.1016/j.chemosphere.2016.08.022>  
49  
50 875 Defarge, N., Spiroux de Vendômois, J., Séralini, G.E., 2018. Toxicity of formulants and heavy  
51 876 metals in glyphosate-based herbicides and other pesticides. *Toxicol. Reports* 5, 156–163.  
52 877 <https://doi.org/10.1016/j.toxrep.2017.12.025>  
53  
54 878 Demetriades, A., Li, X., Ramsey, M.H., Thornton, I., 2010. Chemical speciation and  
55 879 bioaccessibility of lead in surface soil and house dust, Lavrion urban area, Attiki, Hellas.  
56 880 *Environ. Geochem. Health* 32, 529–552. <https://doi.org/10.1007/s10653-010-9315-9>  
57  
58 881 Dong, C., Taylor, M.P., Gulson, B., 2020. A 25-year record of childhood blood lead exposure and  
59 882 its relationship to environmental sources. *Environ. Res.* 186, 109357.  
60 883 <https://doi.org/10.1016/j.envres.2020.109357>  
61  
62  
63  
64  
65



1  
2  
3  
4 884 Dunn, O.J., 1964. Multiple Comparisons Using Rank Sums. *Technometrics* 6, 241–252.  
5 885 Eckert, J.E., 1933. The Flight Range of the Honeybee. *J. Agric. Res.* 47, 257–285.  
6 886 Ellam, R.M., 2010. The graphical presentation of lead isotope data for environmental source  
7 887 apportionment. *Sci. Total Environ.* 408, 3490–3492.  
8 888 <https://doi.org/10.1016/j.scitotenv.2010.03.037>  
9  
10 889 Environmental Protection Agency, U.S., 2017. National Emissions Inventory Data. URL  
11 890 [https://www.epa.gov/air-emissions-inventories/2017-national-emissions-inventory-nei-](https://www.epa.gov/air-emissions-inventories/2017-national-emissions-inventory-nei-data)  
12 891 [data](https://www.epa.gov/air-emissions-inventories/2017-national-emissions-inventory-nei-data) (accessed 8.4.20).  
13  
14 892 Epova, E.N., Bérail, S., Séby, F., Barre, J.P.G., Vacchina, V., Médina, B., Sarthou, L., Donard,  
15 893 O.F.X., 2020. Potential of lead elemental and isotopic signatures for authenticity and  
16 894 geographical origin of Bordeaux wines. *Food Chem.* 303.  
17 895 <https://doi.org/10.1016/j.foodchem.2019.125277>  
18  
19 896 Ferrier, K.L., Perron, J.T., Mukhopadhyay, S., Rosener, M., Stock, J.D., Huppert, K.L., Slosberg,  
20 897 M., 2013. Covariation of climate and long-term erosion rates across a steep rainfall  
21 898 gradient on the Hawaiian island of Kaua’i. *Bull. Geol. Soc. Am.* 125, 1146–1163.  
22 899 <https://doi.org/10.1130/B30726.1>  
23  
24 900 Filippelli, G.M., Laidlaw, M.A.S., 2010. The elephant in the playground: Confronting lead-  
25 901 contaminated soils as an important source of lead burdens to urban populations. *Perspect.*  
26 902 *Biol. Med.* 53, 31–45.  
27  
28 903 Filippelli, G.M., Risch, M., Laidlaw, M.A.S., Nichols, D.E., Crewe, J., 2015. Geochemical legacies  
29 904 and the future health of cities: A tale of two neurotoxins in urban soils. *Elem. Sci. Anthr.* 3,  
30 905 000059. <https://doi.org/10.12952/journal.elementa.000059>  
31  
32 906 Flegal, A.R., Nriagu, J.O., Niemeyer, S., Coale, K.H., 1989. Isotopic tracers of lead contamination  
33 907 in the Great Lakes. *Nature* 339, 455–458. <https://doi.org/10.1038/339455a0>  
34  
35 908 Flegal, A.R., Odigie, K.O., 2020. Distinguishing between natural and industrial lead in consumer  
36 909 products and other environmental matrices. *J. Agric. Food Chem.*  
37 910 <https://doi.org/10.1021/acs.jafc.9b07848>  
38  
39 911 Franklin, J.M., Roscoe, S.M., Loveridge, W.D., Sangster, D.F., 1983. Lead isotope studies in  
40 912 Superior and Southern provinces., Geological Survey of Canada Bulletin.  
41 913 <https://doi.org/10.4095/109256>  
42  
43 914 Galer, S.J.G., Abouchami, W., 1998. Practical Application of Lead Triple Spiking for Correction of  
44 915 Instrumental Mass Discrimination. *Mineral. Mag.* 62A, 491–492.  
45 916 <https://doi.org/10.1180/minmag.1998.62a.1.260>  
46  
47 917 Giglio, A., Ammendola, A., Battistella, S., Naccarato, A., Pallavicini, A., Simeon, E., Tagarelli, A.,  
48 918 Giulianini, P.G., 2017. *Apis mellifera ligustica*, *Spinola 1806* as bioindicator for detecting  
49 919 environmental contamination: a preliminary study of heavy metal pollution in Trieste,  
50 920 Italy. *Environ. Sci. Pollut. Res.* 24, 659–665. <https://doi.org/10.1007/s11356-016-7862-z>  
51  
52 921 Gimeno-García, E., Andreu, V., Boluda, R., 1996. Heavy metals incidence in the application of  
53 922 inorganic fertilizers and pesticides to rice farming soils. *Environ. Pollut.* 92, 19–25.  
54 923 [https://doi.org/10.1016/0269-7491\(95\)00090-9](https://doi.org/10.1016/0269-7491(95)00090-9)  
55  
56 924 Goldhaber, M.B., Church, S.E., Doe, B.R., Aleinikoff, J.N., Brannon, J.C., Podosek, F.A., Mosier,  
57 925 E.L., Taylor, C.D., Gent, C.A., 1995. Lead and sulfur isotope investigation of Paleozoic  
58 926 sedimentary rocks from the southern Midcontinent of the US: implications for  
59 927 paleohydrology and ore genesis of the southeast Missouri lead belts. *Econ. Geol.* 90, 1875–

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
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51  
52  
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57  
58  
59  
60  
61  
62  
63  
64  
65

1910. <https://doi.org/10.2113/gsecongeo.90.7.1875>

Graney, J.R., Halliday, A.N., Keeler, G.J., Nriagu, J.O., Robbins, J.A., Norton, S.A., 1995. Isotopic record of lead pollution in lake sediments from the northeastern United States. *Geochim. Cosmochim. Acta* 59, 1715–1728. <https://doi.org/10.1007/s11270-006-3009-z>

Grousset, F.E., Biscaye, P.E., 2005. Tracing dust sources and transport patterns using Sr, Nd and Pb isotopes. *Chem. Geol.* 222, 149–167. <https://doi.org/10.1016/j.chemgeo.2005.05.006>

Gulson, B.L., 1984. Uranium-lead and lead-lead investigations of minerals from the Broken Hill lodes and mine sequence rocks. *Econ. Geol.* 79, 476–490. <https://doi.org/10.2113/gsecongeo.79.3.476>

Guo, W., Hu, S., Wu, Z., Lan, G., Jin, L., Pang, X., Zhan, J., Chen, B., Tang, Z., 2015. Classification of the geographic origin of cigarettes according to Pb isotope ratios by inductively coupled plasma dynamic reaction cell mass spectrometry. *J. Anal. At. Spectrom.* 30, 986–993. <https://doi.org/10.1039/c4ja00315b>

Han, F.X., Banin, A., Su, Y., Monts, D.L., Plodinec, M.J., Kingery, W.L., Triplett, G.E., 2002. Industrial age anthropogenic inputs of heavy metals into the pedosphere. *Naturwissenschaften* 89, 497–504. <https://doi.org/10.1007/s00114-002-0373-4>

Hanano, D., Weis, D., Scoates, J.S., Aciego, S., DePaolo, D.J., 2010. Horizontal and vertical zoning of heterogeneities in the Hawaiian mantle plume from the geochemistry of consecutive postshield volcano pairs: Kohala-Mahukona and Mauna Kea-Hualalai. *Geochemistry, Geophys. Geosystems* 11, n/a-n/a. <https://doi.org/10.1029/2009gc002782>

Hawaii Climate Portal, National Weather Service, National Oceanic and Atmospheric Administration, USA, 2020. URL [https://www.weather.gov/hfo/rain\\_summary](https://www.weather.gov/hfo/rain_summary) (accessed 8.1.20).

Heinen De Carlo, E., Anthony, S.S., 2002. Spatial and temporal variability of trace element concentrations in an urban subtropical watershed, Honolulu, Hawaii. *Appl. Geochemistry* 17, 475–492. [https://doi.org/10.1016/S0883-2927\(01\)00114-7](https://doi.org/10.1016/S0883-2927(01)00114-7)

Hong, S., Candelone, J., Patterson, C.C., Boutron, C.F., 1994. Greenland Ice Evidence of Hemispheric Lead Pollution Two Millennia Ago by Greek and Roman Civilizations. *Science*. 265, 1841–1843.

Huber, M., Welker, A., Helmreich, B., 2015. Critical review of heavy metal pollution of traffic area runoff: Occurrence, influencing factors, and partitioning. *Sci. Total Environ.* 541, 895–919. <https://doi.org/10.1016/j.scitotenv.2015.09.033>

Jaeger, R.J., Weiss, a L., Manton, W.I., 1998. Isotopic ratio analysis in residential lead-based paint and associated surficial dust. *J. Toxicol. Clin. Toxicol.* 36, 691–703. <https://doi.org/10.3109/15563659809162617>

Jones, C.E., Halliday, A.N., Rea, D.K., Owen, R.M., 2000. Eolian inputs of lead to the North Pacific. *Geochim. Cosmochim. Acta* 64, 1405–1416. [https://doi.org/10.1016/S0016-7037\(99\)00439-1](https://doi.org/10.1016/S0016-7037(99)00439-1)

Kabata-Pendias, A., 2010. *Trace Elements in Soils and Plants*, 4th ed. CRC Press, Taylor & Francis Group, Boca Raton, FL. <https://doi.org/10.1017/CBO9781107415324.004>

Kabata-Pendias, A., 2004. Soil-plant transfer of trace elements - An environmental issue. *Geoderma* 122, 143–149. <https://doi.org/10.1016/j.geoderma.2004.01.004>

Keller, A.T., Regan, L.A., Lundstrom, C.C., Bower, N.W., 2016. Evaluation of the efficacy of spatiotemporal Pb isoscapes for provenancing of human remains. *Forensic Sci. Int.* 261,

1  
2  
3  
4 972 83–92. <https://doi.org/10.1016/j.forsciint.2016.02.006>  
5  
6 973 Klink, K., 1999. Climatological mean and interannual variance of United States surface wind  
7 974 speed, direction and velocity. *Int. J. Climatol.* 19, 471–488.  
8 975 [https://doi.org/10.1002/\(SICI\)1097-0088\(199904\)19:5<471::AID-JOC367>3.0.CO;2-X](https://doi.org/10.1002/(SICI)1097-0088(199904)19:5<471::AID-JOC367>3.0.CO;2-X)  
9  
10 976 Komárek, M., Ettlér, V., Chrastný, V., Mihaljevič, M., 2008. Lead isotopes in environmental  
11 977 sciences: A review. *Environ. Int.* 34, 562–577.  
12 978 <https://doi.org/10.1016/j.envint.2007.10.005>  
13 979 Kristensen, L.J., Taylor, M.P., Evans, A.J., 2016. Tracing changes in atmospheric sources of lead  
14 980 contamination using lead isotopic compositions in Australian red wine. *Chemosphere* 154,  
15 981 40–47. <https://doi.org/10.1016/j.chemosphere.2016.03.023>  
16 982 Kruskal, W.H., Wallis, W.A., 1952. Use of Ranks in One-Criterion Variance Analysis. *J. Am. Stat.*  
17 983 *Assoc.* 47, 583–621.  
18 984 Kurtz, A.C., Derry, L.A., Chadwick, O.A., 2001. Accretion of Asian dust to Hawaiian soils: Isotopic,  
19 985 elemental, and mineral mass balances. *Geochim. Cosmochim. Acta* 65, 1971–1983.  
20 986 [https://doi.org/10.1016/S0016-7037\(01\)00575-0](https://doi.org/10.1016/S0016-7037(01)00575-0)  
21 987 Laidlaw, M.A.S., Filippelli, G.M., 2008. Resuspension of urban soils as a persistent source of lead  
22 988 poisoning in children: A review and new directions. *Appl. Geochemistry* 23, 2021–2039.  
23 989 <https://doi.org/10.1016/j.apgeochem.2008.05.009>  
24 990 Landrigan, P.J., Fuller, R., Acosta, N.J.R., Adeyi, O., Arnold, R., Basu, N., Baldé, A.B., Bertollini, R.,  
25 991 Bose-O'Reilly, S., Boufford, J.I., Breyse, P.N., Chiles, T., Mahidol, C., Coll-Seck, A.M.,  
26 992 Cropper, M.L., Fobil, J., Fuster, V., Greenstone, M., Haines, A., Hanrahan, D., Hunter, D.,  
27 993 Khare, M., Krupnick, A., Lanphear, B., Lohani, B., Martin, K., Mathiasen, K. V., McTeer,  
28 994 M.A., Murray, C.J.L., Ndahimananjara, J.D., Perera, F., Potočník, J., Preker, A.S., Ramesh, J.,  
29 995 Rockström, J., Salinas, C., Samson, L.D., Sandilya, K., Sly, P.D., Smith, K.R., Steiner, A.,  
30 996 Stewart, R.B., Suk, W.A., van Schayck, O.C.P., Yadama, G.N., Yumkella, K., Zhong, M., 2018.  
31 997 The Lancet Commission on pollution and health. *Lancet* 391, 462–512.  
32 998 [https://doi.org/10.1016/S0140-6736\(17\)32345-0](https://doi.org/10.1016/S0140-6736(17)32345-0)  
33 999 Li, I., Cheng, Zhongqi, Paltseva, A., Morin, T., Smith, B., Shaw, R., 2018. Lead in New York city  
34 1000 soils, in: Vasenev, V.I., Dovletyarova, E., Cheng, Z., Valentini, R. (Eds.), *Megacities 2050:*  
35 1001 *Environmental Consequences of Urbanization.* Springer International Publishing AG 2018,  
36 1002 pp. 62–79. [https://doi.org/10.1007/978-3-319-70557-6\\_9](https://doi.org/10.1007/978-3-319-70557-6_9)  
37 1003 Li, M., Weis, D., Smith, K.E., Shiel, A.E., Smith, W.D., Hunt, B.P.V., Torchinsky, A., Pakhomov,  
38 1004 E.A., 2020. Assessing lead sources in fishes of the northeast Pacific Ocean. *Anthropocene*  
39 1005 29. <https://doi.org/10.1016/j.ancene.2019.100234>  
40 1006 Lima, A.L., Bergquist, B.A., Boyle, E.A., Reuer, M.K., Dudas, F.O., Reddy, C.M., Eglinton, T.I.,  
41 1007 2005. High-resolution historical records from Pettaquamscutt River basin sediments: 2. Pb  
42 1008 isotopes reveal a potential new stratigraphic marker. *Geochim. Cosmochim. Acta* 69,  
43 1009 1813–1824. <https://doi.org/10.1016/j.gca.2004.10.008>  
44 1010 Locutura, J., Bel-lan, A., 2011. Systematic urban geochemistry of Madrid, Spain, based on soils  
45 1011 and dust, in: Johnson, C.C., Demetriades, A., Locutura, J., Ottesen, R.T. (Eds.), *Mapping the*  
46 1012 *Chemical Environment of Urban Areas.* John Wiley & Sons, Ltd, pp. 307–347.  
47 1013 Marcantonio, F., Flowers, G., Thien, L., Ellgaard, E., 1998. Lead isotopes in tree rings:  
48 1014 Chronology of pollution in Bayou Trepagnier, Louisiana. *Environ. Sci. Technol.* 32, 2371–  
49 1015 2376. <https://doi.org/10.1021/es980109v>  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

- 1  
2  
3  
4 1016 Marx, S.K., Rashid, S., Stromsoe, N., 2016. Global-scale patterns in anthropogenic Pb  
5 1017 contamination reconstructed from natural archives. *Environ. Pollut.* 213, 283–298.  
6 1018 <https://doi.org/10.1016/j.envpol.2016.02.006>  
7  
8 1019 Masri, S., Kang, C.M., Koutrakis, P., 2015. Composition and sources of fine and coarse particles  
9 1020 collected during 2002–2010 in Boston, MA. *J. Air Waste Manag. Assoc.* 65, 287–297.  
10 1021 <https://doi.org/10.1080/10962247.2014.982307>  
11  
12 1022 Matschullat, J., Ottenstein, R., Reimann, C., 2000. Geochemical background-can we calculate it?  
13 1023 *Environ. Geol.* 39, 990–1000.  
14  
15 1024 McBride, M.B., Shayler, H.A., Spliethoff, H.M., Mitchell, R.G., Marquez-Bravo, L.G., Ferenz, G.S.,  
16 1025 Russell-Anelli, J.M., Casey, L., Bachman, S., 2014. Concentrations of lead, cadmium and  
17 1026 barium in urban garden-grown vegetables: The impact of soil variables. *Environ. Pollut.*  
18 1027 194, 254–261. <https://doi.org/10.1016/j.envpol.2014.07.036>  
19  
20 1028 McConnell, J.R., Wilson, A.I., Stohl, A., Arienzo, M.M., Chellman, N.J., Eckhardt, S., Thompson,  
21 1029 E.M., Pollard, A.M., Steffensen, J.P., 2018. Lead pollution recorded in Greenland ice  
22 1030 indicates European emissions tracked plagues, wars, and imperial expansion during  
23 1031 antiquity. *Proc. Natl. Acad. Sci. U. S. A.* 115, 5726–5731.  
24 1032 <https://doi.org/10.1073/pnas.1721818115>  
25  
26 1033 Medina, B., Augagneur, S., Barbaste, M., Grousset, F.E., Buat-Menard, P., 2000. Influence of  
27 1034 atmospheric pollution on the lead content of wines. *Food Addit. Contam.* 17, 435–445.  
28 1035 <https://doi.org/10.1080/02652030050034019>  
29  
30 1036 Mielke, H.W., 2018. Dynamic geochemistry of tetraethyl lead dust during the 20th century:  
31 1037 Getting the lead in, out, and translational beyond. *Int. J. Environ. Res. Public Health* 15.  
32 1038 <https://doi.org/10.3390/ijerph15050860>  
33  
34 1039 Mielke, H.W., 1994. Lead in New Orleans soils: New images of an urban environment. *Environ.*  
35 1040 *Geochem. Health* 16, 123–128. <https://doi.org/10.1007/BF01747908>  
36  
37 1041 Mielke, H.W., Laidlaw, M.A.S., Gonzales, C.R., 2011. Estimation of leaded (Pb) gasoline's  
38 1042 continuing material and health impacts on 90 US urbanized areas. *Environ. Int.* 37, 248–  
39 1043 257. <https://doi.org/10.1016/j.envint.2010.08.006>  
40  
41 1044 Mielke, H.W., Reagan, P.L., 1998. Soil is an important pathway of human lead exposure.  
42 1045 *Environ. Health Perspect.* 106, 217–229. <https://doi.org/10.1289/ehp.98106s1217>  
43  
44 1046 Millen, T.M., Zartman, R.E., Heyl, A.V., 1995. Pb isotopes from the Upper Mississippi Valley  
45 1047 District: A Regional Perspective, USGS Bulletin 2094-B.  
46  
47 1048 Monastra, V., Derry, L.A., Chadwick, O.A., 2004. Multiple sources of lead in soils from a  
48 1049 Hawaiian chronosequence. *Chem. Geol.* 209, 215–231.  
49 1050 <https://doi.org/10.1016/j.chemgeo.2004.04.027>  
50  
51 1051 Ndung'u, K., Hibdon, S., Véron, A., Flegel, A.R., 2011. Lead isotopes reveal different sources of  
52 1052 lead in balsamic and other vinegars. *Sci. Total Environ.* 409, 2754–2760.  
53 1053 <https://doi.org/10.1016/j.scitotenv.2011.04.001>  
54  
55 1054 Negri, I., Mavris, C., Di Prisco, G., Caprio, E., Pellicchia, M., 2015. Honey bees (*Apis mellifera*, L.)  
56 1055 as active samplers of airborne particulate matter. *PLoS One* 10, 1–22.  
57 1056 <https://doi.org/10.1371/journal.pone.0132491>  
58  
59 1057 Novak, M., Mikova, J., Krachler, M., Kosler, J., Erbanova, L., Prechova, E., Jackova, I., Fottova, D.,  
60 1058 2010. Radial distribution of lead and lead isotopes in stem wood of Norway spruce: A  
61 1059 reliable archive of pollution trends in Central Europe. *Geochim. Cosmochim. Acta* 74,  
62  
63  
64  
65

1  
2  
3  
4 1060 4207–4218. <https://doi.org/10.1016/j.gca.2010.04.059>  
5  
6 1061 NPRI, 2017. Canada's National Pollutant Release Inventory: Data Highlights 2017. Map Facil.  
7 1062 Report. to NPRI 2017, by Ind. Sect. URL [https://www.canada.ca/en/environment-climate-](https://www.canada.ca/en/environment-climate-change/services/national-pollutant-release-inventory/tools-resources-data/fact-sheet.html)  
8 1063 [change/services/national-pollutant-release-inventory/tools-resources-data/fact-](https://www.canada.ca/en/environment-climate-change/services/national-pollutant-release-inventory/tools-resources-data/fact-sheet.html)  
9 1064 [sheet.html](https://www.canada.ca/en/environment-climate-change/services/national-pollutant-release-inventory/tools-resources-data/fact-sheet.html) (accessed 8.4.20).  
10  
11 1065 Nriagu, J.O., 1989. A global assessment of natural sources of atmospheric trace metals. *Nature*  
12 1066 338, 47–49. <https://doi.org/10.1038/338047a0>  
13  
14 1067 O'Connor, D., Hou, D., Ok, Y.S., Lanphear, B.P., 2020. The effects of iniquitous lead exposure on  
15 1068 health. *Nat. Sustain.* 3, 77–79. <https://doi.org/10.1038/s41893-020-0475-z>  
16 1069 O'Connor, D., Hou, D., Ye, J., Zhang, Y., Ok, Y.S., Song, Y., Coulon, F., Peng, T., Tian, L., 2018.  
17 1070 Lead-based paint remains a major public health concern: A critical review of global  
18 1071 production, trade, use, exposure, health risk, and implications. *Environ. Int.* 121, 85–101.  
19 1072 <https://doi.org/10.1016/j.envint.2018.08.052>  
20  
21 1073 Orazio, C., May, T., Gale, R., Meadows, J., Brumbaugh, W., Echols, K., Steiner, W., Berg, C., 2007.  
22 1074 Survey of Chemical Contaminants in the Hanalei River, Kauai, Hawaii, 2001, Scientific  
23 1075 Investigations Report 2007-5096.  
24  
25 1076 Paltseva, A., 2019. Lead and Arsenic Contamination in Urban Soils in New York City.  
26 1077 Patrick, G.J., Farmer, J.G., 2006. A stable lead isotopic investigation of the use of sycamore tree  
27 1078 rings as a historical biomonitor of environmental lead contamination. *Sci. Total Environ.*  
28 1079 362, 278–291. <https://doi.org/10.1016/j.scitotenv.2005.12.004>  
29  
30 1080 Pellicchia, M., Negri, I., 2018. Particulate matter collection by honey bees (*Apis mellifera*, L.)  
31 1081 near to a cement factory in Italy. *PeerJ* 6, e5322. <https://doi.org/10.7717/peerj.5322>  
32  
33 1082 Pellerano, R.G., Uñates, M.A., Cantarelli, M.A., Camiña, J.M., Marchevsky, E.J., 2012. Analysis of  
34 1083 trace elements in multifloral Argentine honeys and their classification according to  
35 1084 provenance. *Food Chem.* 134, 578–582. <https://doi.org/10.1016/j.foodchem.2012.02.125>  
36  
37 1085 Pohl, P., Bielawska-Pohl, A., Dzimitrowicz, A., Jamroz, P., Welna, M., Lesniewicz, A., Szymczycha-  
38 1086 Madeja, A., 2017. Recent achievements in element analysis of bee honeys by atomic and  
39 1087 mass spectrometry methods. *Trends Anal. Chem.* 93, 67–77.  
40 1088 <https://doi.org/10.1016/j.trac.2017.05.009>  
41  
42 1089 Porrini, C., Sabatini, A.G., Girotti, S., Fini, F., Monaco, L., Celli, G., Bortolotti, L., Ghini, S., 2003.  
43 1090 The death of honey bees and environmental pollution by pesticides: The honey bees as  
44 1091 biological indicators. *Bull. Insectology* 56, 147–152.  
45  
46 1092 Pribil, M.J., Maddaloni, M.A., Staiger, K., Wilson, E., Magriples, N., Ali, M., Santella, D., 2014.  
47 1093 Investigation of off-site airborne transport of lead from a superfund removal action site  
48 1094 using lead isotope ratios and concentrations. *Appl. Geochemistry* 41, 89–94.  
49 1095 <https://doi.org/10.1016/j.apgeochem.2013.11.004>  
50  
51 1096 R Core Team, 2020. R: A language and environment for statistical computing. v.4.0.2.  
52 1097 Rauch, J.N., Pacyna, J.M., 2009. Earth's global Ag, Al, Cr, Cu, Fe, Ni, Pb, and Zn cycles. *Global*  
53 1098 *Biogeochem. Cycles* 23. <https://doi.org/10.1029/2008GB003376>  
54  
55 1099 Reimann, C., Flem, B., Fabian, K., Birke, M., Ladenberger, A., Négrel, P., Demetriades, A.,  
56 1100 Hoogewerff, J., Albanese, S., Andersson, M., Arnoldussen, A., Baritz, R., Batista, M.J., Bel-  
57 1101 lan, A., Cicchella, D., Dinelli, E., De Vivo, B., De Vos, W., Duris, M., Dusza-Dobek, A., Eggen,  
58 1102 O.A., Eklund, M., Ernsten, V., Filzmoser, P., Finne, T.E., Flight, D., Forrester, S., Fuchs, M.,  
59 1103 Fugedi, U., Gilucis, A., Gosar, M., Gregorauskiene, V., Gulan, A., Halamić, J., Haslinger, E.,  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 1104 Hayoz, P., Hobiger, G., Hoffmann, R., Hrvatovic, H., Husnjak, S., Janik, L., Johnson, C.C.,  
5 1105 Jordan, G., Kirby, J., Kivisilla, J., Klos, V., Krone, F., Kwecko, P., Kutu, L., Lima, A., Locutura, J.,  
6 1106 Lucivjansky, P., Mackovych, D., Malyuk, B.I., Maquil, R., McLaughlin, M., R.G. Meuli, Miosic,  
7 1107 N., Mol, G., O'Connor, P., Oorts, K., Ottesen, R.T., Pasieczna, A., Petersell, V., Pflaiderer, S.,  
8 1108 Poňavič, M., Prazeres, C., Rauch, U., Salpeteur, I., Schedl, A., Scheib, A., Schoeters, I.,  
9 1109 Sefcik, P., Sellersjö, E., Skopljak, F., Slaninka, I., Šorša, A., Srvkota, R., Stafilov, T., Tarvainen,  
10 1110 T., Trendavilov, V., Valera, P., Verougstraete, V., Vidojević, D., Zissimos, A.M., Zomeni, Z.,  
11 1111 2012. Lead and lead isotopes in agricultural soils of Europe - The continental perspective.  
12 1112 Appl. Geochemistry 27, 532–542. <https://doi.org/10.1016/j.apgeochem.2011.12.012>  
13 1113 Reimann, C., Smith, D.B., Woodruff, L.G., Flem, B., 2011. Pb-concentrations and Pb-isotope  
14 1114 ratios in soils collected along an east-west transect across the United States. Appl.  
15 1115 Geochemistry 26, 1623–1631. <https://doi.org/10.1016/j.apgeochem.2011.04.018>  
16 1116 Rogers, K.M., Sim, M., Stewart, S., Phillips, A., Cooper, J., Douance, C., Pyne, R., Rogers, P., 2014.  
17 1117 Investigating C-4 sugar contamination of manuka honey and other New Zealand honey  
18 1118 varieties using carbon isotopes. J. Agric. Food Chem. 62, 2605–2614.  
19 1119 <https://doi.org/10.1021/jf404766f>  
20 1120 Sangster, D.F., Outridge, P.M., Davis, W.J., 2000. Stable lead isotope characteristics of lead ore  
21 1121 deposits of environmental significance. Environ. Rev. 8, 115–147.  
22 1122 <https://doi.org/10.1139/er-8-2-115>  
23 1123 Sherman, L.S., Blum, J.D., Dvonch, J.T., Gratz, L.E., Landis, M.S., 2015. The use of Pb, Sr, and Hg  
24 1124 isotopes in Great Lakes precipitation as a tool for pollution source attribution. Sci. Total  
25 1125 Environ. 502, 362–374. <https://doi.org/10.1016/j.scitotenv.2014.09.034>  
26 1126 Shiel, A.E., Weis, D., Orians, K.J., 2012. Tracing cadmium, zinc and lead sources in bivalves from  
27 1127 the coasts of western Canada and the USA using isotopes. Geochim. Cosmochim. Acta 76,  
28 1128 175–190. <https://doi.org/10.1016/j.gca.2011.10.005>  
29 1129 Shotyk, W., Weiss, D., Appleby, P.G., Cheburkin, A.K., Frei, R., Gloor, M., Kramers, J.D., Reese, S.,  
30 1130 Van der Knaap, W.O., 1998. History of atmospheric lead deposition since 12,370 14C yr BP  
31 1131 from a peat bog, Jura Mountains, Switzerland. Science. 281, 1635–1640.  
32 1132 Simonetti, A., Gariépy, C., Carignan, J., 2003. Tracing sources of atmospheric pollution in  
33 1133 Western Canada using the Pb isotopic composition and heavy metal abundances of  
34 1134 epiphytic lichens. Atmos. Environ. 37, 2853–2865. [https://doi.org/10.1016/S1352-2310\(03\)00210-3](https://doi.org/10.1016/S1352-2310(03)00210-3)  
35 1135 Simonetti, A., Gariépy, C., Carignan, J., 2000a. Pb and Sr isotopic compositions of snowpack  
36 1136 from Quebec, Canada: Inferences on the sources and deposition budgets of atmospheric  
37 1137 heavy metals. Geochim. Cosmochim. Acta 64, 5–20. [https://doi.org/10.1016/S0016-7037\(99\)00207-0](https://doi.org/10.1016/S0016-7037(99)00207-0)  
38 1138 Simonetti, A., Gariépy, C., Carignan, J., Poissant, L., 2000b. Isotopic evidence of trace metal  
39 1139 sources and transport in eastern Canada as recorded from wet deposition. J. Geophys. Res.  
40 1140 105, 12,263–12,278.  
41 1141 Skorbilowicz, E., Skorbilowicz, M., Cieśluk, I., 2018. Bees as bioindicators of environmental  
42 1142 pollution with metals in an urban area. J. Ecol. Eng. 19, 229–234.  
43 1143 <https://doi.org/10.12911/22998993/85738>  
44 1144 Smith, D., Demetriades, A., de-Caritat, P., Wang, X., 2018. The history, progress, and future of  
45 1145 global-scale geochemical mapping. Geochim. Bras. 32, 115–135.  
46 1146  
47 1147  
48 1148  
49 1149  
50 1150  
51 1151  
52 1152  
53 1153  
54 1154  
55 1155  
56 1156  
57 1157  
58 1158  
59 1159  
60 1160  
61  
62  
63  
64  
65

1  
2  
3  
4 1148 <https://doi.org/10.21715/gb2358-2812.2018322115>  
5  
6 1149 Smith, K.E., Weis, D., 2020. Evaluating Spatiotemporal Resolution of Trace Element  
7 1150 Concentrations and Pb Isotopic Compositions of Honeybees and Hive Products as  
8 1151 Biomonitoring for Urban Metal Distribution. *GeoHealth* 4.  
9 1152 <https://doi.org/10.1029/2020GH000264>  
10  
11 1153 Smith, K.E., Weis, D., Amini, M., Shiel, A.E., Lai, V.W.-M., Gordon, K., 2019. Honey as a  
12 1154 biomonitor for a changing world. *Nat. Sustain.* 2, 223–232.  
13 1155 <https://doi.org/10.1038/s41893-019-0243-0>  
14  
15 1156 Smith, K.E., Weis, D., Chauvel, C., Moulin, S., 2020. Honey Maps the Pb Fallout from the 2019  
16 1157 Fire at Notre-Dame Cathedral, Paris: A Geochemical Perspective. *Environ. Sci. Technol.*  
17 1158 *Lett.* <https://doi.org/10.1021/acs.estlett.0c00485>  
18  
19 1159 Solayman, M., Islam, M.A., Paul, S., Ali, Y., Khalil, M.I., Alam, N., Gan, S.H., 2016.  
20 1160 Physicochemical Properties, Minerals, Trace Elements, and Heavy Metals in Honey of  
21 1161 Different Origins: A Comprehensive Review. *Compr. Rev. Food Sci. Food Saf.* 15, 219–233.  
22 1162 <https://doi.org/10.1111/1541-4337.12182>  
23  
24 1163 Spencer, K.J., De Carlo, E.H., McMurtry, G.M., 1995. Isotopic clues to sources of natural and  
25 1164 anthropogenic lead in sediments and soils from Oahu, Hawaii. *Pacific Sci.* 49, 492–510.  
26 1165 Spengler, S.R., Heskett, M.D., Gray, J.I., 2019. Pesticide levels in streams and sediments on the  
27 1166 islands of Oahu and Kauai, Hawaii. *Int. J. Environ. Impacts Manag. Mitig. Recover.* 2, 283–  
28 1167 299. <https://doi.org/10.2495/ei-v2-n3-283-299>  
29  
30 1168 Sturges, W., Barrie, L., 1987. Lead 206/207 isotope ratios in the atmosphere of North America  
31 1169 as tracers of US and Canadian emissions. *Nature* 329, 144–146.  
32  
33 1170 Sutherland, R.A., Day, J.P., Bussen, J.O., 2003. Lead concentrations, isotope ratios, and source  
34 1171 apportionment in road deposited sediments, Honolulu, Oahu, Hawaii. *Water. Air. Soil*  
35 1172 *Pollut.* 142, 165–186. <https://doi.org/10.1023/A:1022026612922>  
36  
37 1173 Taylor, M.P., 2019. Bees as biomarkers. *Nat. Sustain.* 2, 169–170.  
38 1174 <https://doi.org/10.1038/s41893-019-0247-9>  
39  
40 1175 USEPA, 1999. Background report on fertilizer use, contaminants and regulations: EPA 747-R-98-  
41 1176 003. Washington, DC.  
42 1177 USGS, 2020. U.S. Geological Survey, National Water Information System: Web Interface. URL  
43 1178 <https://waterdata.usgs.gov/usa/nwis/uv?16031000> (accessed 10.21.20).  
44  
45 1179 Van der Steen, J.J.M., de Kraker, J., Grotenhuis, T., 2015. Assessment of the Potential of  
46 1180 Honeybees (*Apis mellifera* L.) in Biomonitoring of Air Pollution by Cadmium, Lead and  
47 1181 Vanadium. *J. Environ. Prot. (Irvine, Calif.)* 06, 96–102.  
48 1182 <https://doi.org/10.4236/jep.2015.62011>  
49  
50 1183 Voerkelius, S., Lorenz, G.D., Rummel, S., Quénel, C.R., Heiss, G., Baxter, M., Brach-Papa, C.,  
51 1184 Deters-Iltzelsberger, P., Hoelzl, S., Hoogewerff, J., Ponzevera, E., Van Bockstaele, M.,  
52 1185 Ueckermann, H., 2010. Strontium isotopic signatures of natural mineral waters, the  
53 1186 reference to a simple geological map and its potential for authentication of food. *Food*  
54 1187 *Chem.* 118, 933–940. <https://doi.org/10.1016/j.foodchem.2009.04.125>  
55  
56 1188 Von Uexküll, O., Skerfving, S., Doyle, R., Braungart, M., 2005. Antimony in brake pads—a  
57 1189 carcinogenic component? *J. Clean. Prod.* 13, 19–31.  
58 1190 <https://doi.org/10.1016/j.jclepro.2003.10.008>  
59  
60 1191 Weis, D., Garcia, M.O., Rhodes, J.M., Jellinek, M., Scoates, J.S., 2011. Role of the deep mantle in

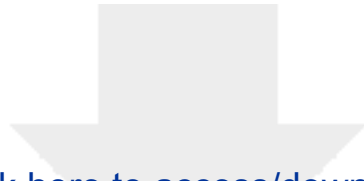
1  
2  
3  
4 1192 generating the compositional asymmetry of the Hawaiian mantle plume. *Nat. Geosci.* 4,  
5 831–838. <https://doi.org/10.1038/ngeo1328>  
6 1193  
7 1194 Weiss, A.L., Caravanos, J., Blaise, M.J., Jaeger, R.J., 2006. Distribution of lead in urban roadway  
8 1195 grit and its association with elevated steel structures. *Chemosphere* 65, 1762–1771.  
9 <https://doi.org/10.1016/j.chemosphere.2006.04.079>  
10 1196  
11 1197 Weiss, D., Shotyk, W., Kempf, O., 1999. Archives of atmospheric lead pollution.  
12 1198 *Naturwissenschaften* 86, 262–275. <https://doi.org/10.1007/s001140050612>  
13 1199  
14 1200 Williamson, N.M.B., Weis, D., Scoates, J.S., Pelletier, H., Garcia, M.O., 2019. Tracking the  
15 1201 Geochemical Transition Between the Kea-Dominated Northwest Hawaiian Ridge and the  
16 1202 Bilateral Loa-Kea Trends of the Hawaiian Islands. *Geochemistry, Geophys. Geosystems* 20,  
17 4354–4369. <https://doi.org/10.1029/2019GC008451>  
18 1203  
19 1204 Wilson, J.G., Kingham, S., Pearce, J., Sturman, A.P., 2005. A review of intraurban variations in  
20 1205 particulate air pollution: Implications for epidemiological research. *Atmos. Environ.* 39,  
21 6444–6462. <https://doi.org/10.1016/j.atmosenv.2005.07.030>  
22 1206  
23 1207 Wu, L., Du, B., Vander Heyden, Y., Chen, L., Zhao, L., Wang, M., Xue, X., 2017. Recent  
24 1208 advancements in detecting sugar-based adulterants in honey – A challenge. *TrAC - Trends*  
25 1209 *Anal. Chem.* 86, 25–38. <https://doi.org/10.1016/j.trac.2016.10.013>  
26 1210  
27 1211 Zahran, S., Mielke, H.W., McElmurry, S.P., Filippelli, G.M., Laidlaw, M.A.S., Taylor, M.P., 2013.  
28 1212 Determining the relative importance of soil sample locations to predict risk of child lead  
29 1213 exposure. *Environ. Int.* 60, 7–14. <https://doi.org/10.1016/j.envint.2013.07.004>  
30 1214  
31 1215 Zarić, N.M., Ilijević, K., Stanisavljević, L., Gržetić, I., 2016. Metal concentrations around thermal  
32 1216 power plants, rural and urban areas using honeybees (*Apis mellifera* L.) as bioindicators.  
33 1217 *Int. J. Environ. Sci. Technol.* 13, 413–422. <https://doi.org/10.1007/s13762-015-0895-x>  
34 1218  
35 1219 Zhao, M., Zhang, Y., Ma, W., Fu, Q., Yang, X., Li, C., Zhou, B., Yu, Q., Chen, L., 2013.  
36 1220 Characteristics and ship traffic source identification of air pollutants in China’s largest port.  
37 1221 *Atmos. Environ.* 64, 277–286. <https://doi.org/10.1016/j.atmosenv.2012.10.007>  
38 1222  
39 1223 Zhou, X., Taylor, M.P., Davies, P.J., 2018a. Tracing natural and industrial contamination and lead  
40 1224 isotopic compositions in an Australian native bee species. *Environ. Pollut.* 242, 54–62.  
41 <https://doi.org/10.1016/j.envpol.2018.06.063>  
42 1225  
43 1226 Zhou, X., Taylor, M.P., Davies, P.J., Prasad, S., 2018b. Identifying sources of environmental  
44 1227 contamination in European honey bees (*Apis mellifera*) using trace elements and lead  
45 1228 isotopic compositions. *Environ. Sci. Technol.* 52, 991–1001.  
46 <https://doi.org/10.1021/acs.est.7b04084>  
47 1229  
48 1230 Zhou, X., Taylor, M.P., Salouros, H., Prasad, S., 2018c. Authenticity and geographic origin of  
49 1231 global honeys determined using carbon isotope ratios and trace elements. *Sci. Rep.* 8, 1–  
50 11. <https://doi.org/10.1038/s41598-018-32764-w>  
51 1232





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**Regional and global perspectives of honey as a record of lead in the environment**

Kate E. Smith, Dominique Weis, Sean R. Scott, Carl J. Berg, Yaffa Segal, Philippe Claeys

**CONFLICT OF INTEREST STATEMENT**

The authors declare no conflicts of interest relevant to this study.

**Declaration of interests**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: